# An assessment of changes in the daytime recreational fishery of Lake Macquarie following the establishment of a 'Recreational Fishing Haven' 

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A full list of all personnel that were involved with the 1999/2000 and/or 2003/2004 surveys and their affiliations is provided in Section 10 of this report.

We would like to thank all of the recreational fishers that participated in the surveys. The successful completion of this work was made possible by their continual co-operation and support.

This report is dedicated to the memories of our colleague Doug Chapman and to Bill Gray, a driving force for a net free Lake Macquarie. Their energy and dedication will be missed by all.

## EXECUTIVE SUMMARY

Resource sharing and the allocation of fisheries resources between recreational and commercial user groups has long been a contentious management issue in New South Wales (NSW). The introduction of a general recreational fishing fee in March 2001 generated considerable funding that was used to undertake significant changes in the management of fisheries in NSW. Lake Macquarie was zoned a 'Recreational Fishing Haven' (RFH) following extensive community consultation. This management initiative changed the allocation of fisheries resources in this waterway between the recreational and commercial sectors. This major re-allocation of access to the estuarine fisheries resources in Lake Macquarie has undoubtedly created additional recreational fishing opportunities. Thus, there was an important need to assess whether the recreational fisheries in this RFH were improving and providing better quality recreational fishing. This report focuses on comparisons made between two separate daytime recreational fishing surveys of Lake Macquarie (including Swansea Channel). The first annual survey was done during the pre-RFH period (March 1999 to February 2000) and the second annual survey was done during the postRFH period (December 2003 to November 2004). These annual surveys provide a snapshot of the recreational fishery of Lake Macquarie before RFH implementation and after RFH implementation. The same complemented, on-site, survey design was used in both surveys. The shore-based fishery was assessed by using a roving(effort)-roving(harvest) design combination and the boat-based fishery was assessed by using a roving(effort)-access(harvest) design combination using stratified random sampling methods.

The two recreational fishing surveys provide evidence of a relatively productive recreational fishery in Lake Macquarie. Comparisons made between the two separate daytime recreational fishing surveys indicate that the post-RFH recreational fishery was very different to the fishery that had existed prior to the implementation of the RFH. We documented statistically significant increases in recreational harvest for some prized recreational species and also some significant decreases for some other important recreational species. Overall, the indicators of recreational fishing quality that we examined indicated that the post-RFH fishery had improved in many ways since the pre-RFH survey period. A summary of the evidence provided in this report is that:
(a) the recreational harvest in both survey years was dominated by a relatively small number of taxa, however, the composition and relative contribution of these dominant taxa changed markedly between survey years. These changes occurred even though there was no significant difference between survey years in the total annual harvest, by number or weight, for the whole fishery;
(b) the recreational harvest of dusky flathead, tailor, sand whiting and trumpeter whiting (number and weight) and large-toothed flounder (weight only) increased significantly during the post-RFH survey year;
(c) the recreational harvest of common squid, yellow-finned leatherjacket and sand mullet, by number and weight, decreased significantly during the post-RFH survey year;
(d) total fishing effort (boat and shore combined) showed little change (about 2\%), however, different trends were evident in the boat-based and shore-based fisheries. Fishing effort in the larger boat-based fishery increased by about $13 \%$ but this change was not statistically significant. In contrast, there was a statistically significant reduction of about $22 \%$ in the level of shore-based fishing;
(e) seasonal harvest rate comparisons between survey years tended to confirm the increasing or decreasing trends found in the annual recreational harvest estimates for the main species;
(f) comparisons of length frequency information, mean and median lengths between survey years indicated that most species were harvested at larger sizes during the post-RFH survey year. The mean and median sizes of dusky flathead, sand whiting, tailor, common squid, yellowfin bream, blue swimmer crab, large-toothed flounder and sand mullet were all larger during the second survey year.

This survey provides the first snapshot (point estimate) of the Lake Macquarie recreational fishery following the establishment of the waterway as a RFH. On-site surveys of recreational fishing are valuable tools for collecting information to describe the status of a fishery and any changes that may have occurred since previous survey periods. On-site surveys of the recreational fishery should be repeated regularly (every 3-5 years) to monitor the recreational fishery in Lake Macquarie.

## 1. INTRODUCTION

Resource sharing and the allocation of fisheries resources between recreational and commercial user groups has long been a contentious management issue in New South Wales (e.g. NSW Parliament - Fisheries Inquiry Commission 1880). Since the Fisheries Inquiry of 1880, the recreational sector has continued to grow and this has led to increased conflict with the commercial sector as both groups strive to maximise their share of limited fisheries resources. An extensive investigation into the hydrology, geology and ecology of Lake Macquarie was initiated in 1955 as a result of emphatic allegations that the fish stocks in Lake Macquarie had been depleted by commercial over-exploitation (see Baas Becking 1959; Baas Becking et al. 1959; Davis 1959; MacIntyre 1959; Spencer 1959; Thomson 1959a, 1959b, 1959c, 1959d, \& 1959e; Wood 1959a \& 1959b). Similar allocation disputes within NSW have been concentrated in other estuarine fisheries, usually near large metropolitan areas such as Botany Bay and Sydney Harbour (Ruello and Henry 1977, State Pollution Control Commission (SPCC) 1981, Henry 1984).

Over the past 100 years, the ecosystem of Lake Macquarie has been placed under increasing stress by the combined effects of heavy industry, coal mining, the construction and operation of power stations, commercial and recreational fishing, tourism, non-extractive recreational usage and a variety of agricultural and urban land uses within the catchment system of the Lake (SPCC 1983, Lake Macquarie Taskforce 1999). These stresses have had negative impacts on the Lake Macquarie ecosystem and on the amenity and quality of the recreational and commercial fisheries of Lake Macquarie (Lake Macquarie Taskforce 1999). In response to public concerns about these issues, the NSW government established a taskforce in 1998 to address the issues affecting the health of Lake Macquarie. The Lake Macquarie Taskforce (1999) report provided a comprehensive assessment and integrated plan for improving the conditions of Lake Macquarie and its catchment. The report documented the conflicting views of local commercial and recreational fishing groups about the status of the fisheries resources within the Lake and their preferred management options (Lake Macquarie Taskforce 1999). Thus, there was a need to collect quantitative information to describe the recreational fishery of Lake Macquarie and to compare the relative size of commercial and recreational harvests. A survey of daytime recreational fishing, commenced in March 1999 and completed at the end of February 2000, provided the baseline quantitative information needed to describe and assess the status of the recreational fishery in Lake Macquarie (Steffe and Chapman 2003).

The introduction of a general recreational fishing fee in March 2001 generated funding that was used to undertake significant changes in the management of fisheries in NSW. Extensive community consultation was undertaken to identify suitable estuarine areas that could be zoned 'Recreational Fishing Havens' (RFH). The intent was that areas declared 'Recreational Fishing Havens' would improve recreational fishing opportunities by removing commercial fishing from them. Thirty locations, including the whole of Lake Macquarie, were declared 'Recreational Fishing Havens' during the period May to September 2002. This resulted in a total estuarine area of $27 \%$ being made substantially free of commercial fishing (some RFH areas still have limited commercial fishing). This major re-allocation of access to the estuarine fisheries resources in NSW has undoubtedly created additional recreational fishing opportunities. Thus, there was an important need to assess whether these 'Recreational Fishing Havens' were actually improving the recreational fisheries.

The previous recreational fishing survey done in Lake Macquarie during 1999-2000 (Steffe and Chapman 2003) provided a pre-RFH benchmark that could be used to assess any post-RFH changes that had occurred in the fishery. Hence, another survey of recreational fishing was done so
that we could assess changes in the harvest, effort and quality of fishing that had occurred after the implementation of the RFH.

## 2. OBJECTIVES

The principal aims of this project were:

- To estimate the level of daytime recreational fishing effort and harvest in Lake Macquarie during the annual period, December 2003 to November 2004 inclusive.
- To assess changes in the fishing effort and harvest of recreational fishers that had occurred since the establishment of Lake Macquarie as a Recreational Fishing Haven (RFH) in May 2002.
- To use selected indicators of recreational fishing quality to assess changes in the Lake Macquarie fishery after its establishment as a Recreational Fishing Haven.


## 3. METHODS

### 3.1. General comments

Data comparisons are derived from two separate recreational fishing surveys of Lake Macquarie (including the Swansea Channel). The first annual survey was carried out during March 1999 to February 2000 inclusive and represents a snapshot of the recreational fishery (boat and shore) before the area was declared a Recreational Fishing Haven. The second annual survey was carried out during December 2003 to November 2004 inclusive and represents a snapshot of the recreational fishery covering a period of 1.5 to 2.5 years after the area was made a Recreational Fishing Haven. The same survey methods were used for both survey years, however, the level of daily replication was greater in the second survey period (see Table 1). Description of study area and access for recreational fishers to the fishery.

An error in the structure of a database query was found during the preparation of comparative data analyses from the two survey periods. This error resulted in the overestimation of the harvest (number of fish and weight) reported for the first survey period by Steffe and Chapman (2003). All fishing effort and harvest rate estimates reported by Steffe and Chapman (2003) are correct. Revised harvest estimates for the first survey period are presented in this report.

### 3.2. Description of study area and access for recreational fishers to the fishery

Lake Macquarie $\left(33^{0} 03\right.$ ' $\left.\mathrm{S} 151^{0} 36^{\prime} \mathrm{E}\right)$ is a large coastal lagoon situated to the south of the industrial city of Newcastle, the second largest coastal centre in New South Wales (NSW) on the east coast of Australia (Fig. 1). Lake Macquarie has a surface area of about $115 \mathrm{~km}^{2}$, a total catchment area of about $700 \mathrm{~km}^{2}$ and an average depth of about 6 to 7 meters (Baas Becking et al. 1959, Roy et al. 2001). The Lake is a wave-dominated barrier estuary (Roy et al. 2001) which is connected to the ocean by a permanently-open channel at Swansea. The Swansea Channel is a relatively narrow, shallow area that is characterised by strong tidal currents. The relatively small size of the Swansea Channel in comparison to Lake Macquarie makes it a barrier that restricts tidal exchange between the Lake and the ocean to about $1 \%$ of the Lake volume each tidal cycle (Spencer 1959). The semidiurnal tidal range in NSW coastal waters is about 2.0 meters but within the Lake itself the tidal range averages only 6 millimeters (SPCC 1983). Lake Macquarie contains approximately $1.0 \mathrm{~km}^{2}$ of mangroves, approximately $13.4 \mathrm{~km}^{2}$ of seagrass and approximately $0.7 \mathrm{~km}^{2}$ of saltmarsh vegetation (Roy et al. 2001).

The fisheries resources within Lake Macquarie and the Swansea Channel were accessible to recreational fishers from boats and from the shore. Boat-based fishers were able to access the recreational fishery from a great number of access points (Figure 1, Appendix 1). During the survey periods, there were 38 public boat ramps, more than 2100 boat moorings, a multitude of private homes located on the edge of the Lake from which small boats could be launched and 9 caravan parks located near the Lake (some of these had their own boat ramps). Shoreline access to the recreational fishery was diffuse, even though there were large areas of shoreline which were not very accessible because of the vegetation, topography or restrictions to public access. There were 31 public wharves and jetties (Lake Macquarie Taskforce 1999) and there were about 1060 private jetties in 1998 (Central Mapping Authority - Department of Land and Water Conservation) throughout the Lake (Fig.1). There were two recognised camping grounds and all 9 caravan parks around the Lake provided access to the shoreline. The two outlets which discharge heated water from the power stations and adjacent areas within the thermal plumes were popular recreational fishing areas and were accessible to shore-based anglers during the first survey period.


Figure 1. Map of Lake Macquarie showing the spatial extent of the survey and the boundaries used to divide the fishery into the Northern Lake, Southern Lake and Swansea Channel areas. Circled numbers refer to recognised boat ramps, see Appendix 1 for location descriptions.

Sample sizes（number of days spent interviewing and the number of replicate progressive counts of effort），number of interviews，number of refusals and refusal rates for the boat－based and shore－based recreational fisheries in Lake Macquarie for each annual survey period． \begin{tabular}{ccc}
\multicolumn{3}{c}{ SHORE FISHERY } <br>
\hline $\begin{array}{c}\text { Number of } \\
\text { Interviews }\end{array}$ \& $\begin{array}{c}\text { Number of } \\
\text { Refusals }\end{array}$ \& $\begin{array}{c}\text { Refusal } \\
\text { Rate（\％）}\end{array}$ <br>
\hline 154 \& 4 \& $2.6 \%$

 

154 \& 4 \& $2.6 \%$ <br>
201 \& 3 \& $1.5 \%$ <br>
$\mathbf{3 5 5}$ \& $\mathbf{7}$ \& $2.0 \%$ <br>
122 \& 4 \& $3.3 \%$ <br>
161 \& 0 \& $0.0 \%$ <br>
$\mathbf{2 8 3}$ \& $\mathbf{4}$ \& $1.4 \%$ <br>
149 \& 1 \& $0.7 \%$ <br>
262 \& 1 \& $0.4 \%$ <br>
$\mathbf{4 1 1}$ \& $\mathbf{2}$ \& $0.5 \%$ <br>
205 \& 2 \& $1.0 \%$ <br>
208 \& 4 \& $1.9 \%$ <br>
$\mathbf{4 1 3}$ \& $\mathbf{6}$ \& $1.5 \%$ <br>
\hline 630 \& 11 \& $1.7 \%$ <br>
832 \& 8 \& $1.0 \%$ <br>
$\mathbf{1 4 6 2}$ \& $\mathbf{1 9}$ \& $1.3 \%$ <br>
\& \& <br>
\& \&
\end{tabular}

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|  | のヘットナコかナさ |
|  |  |




| SURVEY YEAR 1 | No．Days | Effort | Interview |
| :--- | :--- | :--- | :--- |

No．Days Effort Interview

 \begin{tabular}{llrrr}
\& Weekend \& $\mathbf{2 9}$ \& $\mathbf{1 0}$ \& $\mathbf{8}$ <br>
\& Total \& $\mathbf{9 2}$ \& $\mathbf{1 0}$ \& 4 <br>
Winter 1999 \& Weekday \& 65 \& 6 \& 4 <br>
\& Weekend \& 27 \& 6 \& 4 <br>
\& Total \& $\mathbf{9 2}$ \& $\mathbf{1 2}$ \& $\mathbf{8}$ <br>
Spring 1999 \& Weekday \& 64 \& 6 \& 6 <br>
\& Weekend \& 27 \& 6 \& 6 <br>
\& Total \& $\mathbf{9 1}$ \& $\mathbf{1 2}$ \& $\mathbf{1 2}$ <br>
Summer \& Weekday \& 60 \& 6 \& 7 <br>
$1999 / 2000$ \& Weekend \& 31 \& 6 \& 6 <br>
\& Total \& $\mathbf{9 1}$ \& $\mathbf{1 2}$ \& $\mathbf{1 3}$ <br>
\hline Survey Year 1 \& Weekday \& 252 \& 23 \& 21 <br>
Total \& Weekend \& 114 \& 23 \& 20 <br>
\& Total \& $\mathbf{3 6 6}$ \& $\mathbf{4 6}$ \& $\mathbf{4 1}$

 

\multicolumn{5}{l}{ SURVEY YEAR 2} <br>
\hline Summer \& Weekday \& 61 \& 10 \& 9 <br>
$2003 / 2004$ \& Weekend \& 30 \& 10 \& 9 <br>
\& Total \& $\mathbf{9 1}$ \& $\mathbf{2 0}$ \& $\mathbf{1 8}$ <br>
Autumn 2004 \& Weekday \& 63 \& 10 \& 9 <br>
\& Weekend \& 29 \& 10 \& 9 <br>
\& Total \& $\mathbf{9 2}$ \& $\mathbf{2 0}$ \& $\mathbf{1 8}$ <br>
Winter 2004 \& Weekday \& 65 \& 10 \& 9 <br>
\& Weekend \& 27 \& 10 \& 9 <br>
\& Total \& $\mathbf{9 2}$ \& $\mathbf{2 0}$ \& $\mathbf{1 8}$ <br>
Spring 2004 \& Weekday \& 64 \& 10 \& 9 <br>
\& Weekend \& 27 \& 10 \& 9 <br>
\& Total \& $\mathbf{9 1}$ \& $\mathbf{2 0}$ \& $\mathbf{1 8}$ <br>
\hline Survey Year 2 \& Weekday \& 253 \& 40 \& 36 <br>
\& Total \& Weekend \& 113 \& 40 <br>
\& Total \& $\mathbf{3 6 6}$ \& $\mathbf{8 0}$ \& $\mathbf{7 2}$ <br>
\hline
\end{tabular}

## Table 1.

Recreational anglers were excluded from the immediate vicinity of the hot water outlet at the Eraring power station during the second survey period. This area had previously received high levels of shore-based angling effort, particularly during Winter and Spring. Most of the Swansea Channel was accessible to shore-based anglers. The channel has long areas of breakwall and some public jetties that were all frequently used as fishing platforms by anglers.

### 3.3. Survey design

The same complemented, on-site, survey design (see Pollock et al. 1994 for a review of angler survey methods) was used to assess the recreational fisheries prior to and after the implementation of the Recreational Fishing Haven in Lake Macquarie. The shore-based fishery was assessed by using a roving(effort)-roving(harvest) design combination whereas the boat-based fishery was assessed by using a roving(effort)-access(harvest) design combination. Stratified random sampling methods were used with days being the primary sampling unit for all strata. By definition, a survey day started at sunrise and ended at sunset.

### 3.3.1. Spatial and temporal sampling frames and stratification

The spatial sampling frame (geographical boundary) of the two recreational fishing surveys is shown in Figure 1. Lake Macquarie (Fig. 1) was stratified into three distinct areas: (a) the Swansea Channel area; (b) the Northern Lake area; and (c) the Southern Lake area; to improve the precision of the estimates of effort and harvest for the total fishery. These spatial strata were selected to reflect major differences in fish habitats, commercial fishing practices (prior to the creation of the RFH ) and perceived differences in recreational fishing quality among the three areas.

The temporal sampling frame of each survey spanned a one year period. Each survey year was stratified into seasons and day-types within season (Weekdays and Weekend days). Public holidays were classified as weekend days. The sequence of seasonal sampling differed between survey years. Survey work done during the first survey year covered Autumn, Winter, Spring, and then Summer, whereas, the sequence of surveying during the second survey year was Summer, Autumn, Winter and then Spring. This difference in the sequence of seasonal sampling is important when considering seasonal comparisons between survey periods because the Summer season comparisons are based on a four year difference whilst the Autumn, Spring and Winter seasonal comparisons are based on a five year difference between sampling periods.

### 3.4. Data collection for the boat-based and shore-based recreational fisheries

Two independent datasets were collected and used to estimate recreational fishing effort and harvest rates. These datasets consisted of: (a) progressive counts of recreational fishing effort; and (b) interviews with recreational fishing parties. These two datasets were then used to obtain estimates of boat-based and shore-based recreational harvest.

### 3.4.1. Progressive counts of recreational fishing effort

Estimates of recreational fishing effort for the boat-based fishery and the shore-based fishery were made with progressive counts on randomly selected survey days. Progressive counts were made separately of all boats and all shore-based persons that were observed to be involved in some type of recreational fishing activity. These recreational fishing activities included all forms of angling and the setting, checking and retrieval of crab nets, but excluded activities such as spearfishing, bait collecting and prawning. We specifically excluded boats traveling across the estuary and anglers moving along the shore from the counts (even when recreational fishing gear was visible) when it was not possible to determine their destination or their immediate intent to engage in any recreational fishing activity. All boats engaged in drift fishing were included in the counts when
they were observed traveling to start another "drift" upstream. Drift fishing was common throughout Swansea Channel and the Southern and Northern Lake areas.

The time needed to complete a circuit of the entire Lake Macquarie fishery by boat was determined during a pilot study. These data were used to schedule the starting times for the progressive counts by picking one of a set of discrete possible starting times as recommended by Hoenig et al. (1993). The starting location and direction of travel were randomly selected for each scheduled progressive count. This progressive count method will, under very general conditions, provide unbiased estimates of fishing effort during the day (Hoenig et al. 1993). The collection of recreational effort data by means of these progressive counts was, in a statistical sense, independent of the collection of interview data. The number of replicate progressive counts done during each survey year for each area of the Lake, and for each day-type stratum within each season is summarised in Table 1. The level of replication achieved during survey years 1 and 2 respectively represents annual sampling fractions of about $20 \%$ and $35 \%$ for the weekend day-type stratum and about $9 \%$ and $16 \%$ for the weekday stratum (Table 1).

### 3.4.2. Interviews with recreational fishing parties

Fishing parties were approached and asked to participate in the survey by providing information about their fishing trip and their harvest. Attempts were made to interview all recreational fishing parties encountered (shore-based and boat-based), however, during periods of high recreational activity it was necessary to systematically subsample every second or third fishing party (depending on the number of fishing parties available for interview). The number of days spent interviewing recreational fishing parties and the number of interviews obtained are summarised for each day-type within each season for both survey years (Table 1). Refusals to provide information, or to show the fish retained, were also recorded (Table 1). We asked co-operative recreational fishers about their targeting preferences during their current fishing trip, the time they started fishing and their fishing locations. We also recorded the number of fishers in the fishing party (non-fishers were not included as part of a fishing party). The retained catch was identified by field staff and, whenever possible, measurements of all fish (fork length), crabs (carapace length) and squid (mantle length) were taken to the nearest whole centimetre. When fishers were in a hurry to leave the ramp and it was not possible to measure all fish, crabs and squid, the survey personnel were instructed to record counts of the identified harvest and attempt to measure a sub-sample of the harvest. Machine-readable interview forms were used to record the information from interviews.

Most interviews with recreational fishing parties were done in areas that had public access. Therefore, we assumed that the fishing activities of recreational fishers using the public boat ramps, wharves and easily accessible shoreline were representative of recreational fishing parties that used private access points to enter and leave the fishery. Although we did not formally test this important assumption, we have no reason to expect that fishers using private access points and other upstream boat ramps would have behaved differently to those fishers that used the public boat ramps within the survey area because these populations of fishers (regardless of where they entered the fishery) use the same methods to target the same species in the same fishing grounds within the survey area. Some limited interview data from fishers using private access points was collected during the surveys. A graphical comparison of these private access point data and the public access point data did not show any major differences between them.

### 3.5. Estimation methods

The following sections provide brief explanations of the estimation methods used to calculate: (a) fishing effort; (b) harvest rates for the boat-based and shore-based fisheries; and (c) harvest. Equations for estimating total recreational fishing effort, recreational harvest rates, and total
recreational harvest for the boat-based and shore-based fisheries are provided by Steffe and Chapman (2003). Detailed explanations of the statistical procedures used can be found in Cochran (1953), Robson (1960, 1961 \& 1991), Yates (1965), Malvestuto (1983), Hayne (1991), Hoenig et al. (1993 \& 1997) and Pollock et al. (1994 \& 1997).

### 3.5.1. Effort estimation for the boat-based and shore-based recreational fisheries

Estimation of recreational effort was done separately for the boat-based fishery (units of boat hours) and the shore-based fishery (units of fisher hours). The base level of effort estimation was a day-type stratum within a season for each of the three areas in the Lake Macquarie fishery (see Figure 1). The progressive counts of recreational fishing boats and shore-based fishers were expanded separately to estimate the daily effort for each fishing day that was sampled. These daily effort replicates for each area were expanded to estimate day-type stratum totals within each season. Seasonal estimates of effort were obtained by adding the estimates from the day-type strata together. Total fishery effort estimates (boat-based plus shore-based) were calculated after the conversion of the boat-based data into units of fisher hours. Whenever strata were combined their variances were additive.

### 3.5.2. Harvest rate estimators for the boat-based and shore-based fisheries

Boat-based fishing parties were approached at boat ramps when they returned from their fishing trip. The harvest rate information collected during these access point interviews is based on completed trips (Malvestuto 1983, Hayne 1991, Pollock et al. 1994, Pollock et al. 1994 \& 1997). When the objective is to estimate total harvest, and the interview data are derived from completed trips, the correct harvest rate estimator to use is the 'ratio of means' (Jones et al. 1995, Pollock et al. 1997). This estimator is essentially the ratio of mean harvest to mean effort on a given day. The mean daily 'ratio of means' estimator calculated for each base stratum was used for estimating the harvest of the boat-based fishery.

The diffuse access across large stretches of shoreline and breakwater compelled us to use roving survey methods to locate shore-based fishers. The shore-based fishery within the survey area was searched entirely at least once (usually many times) during each survey day by interviewers, thus providing coverage of the entire shore-based fishery on each survey day. Shore-based fishing parties were approached during their fishing trips by field staff. Therefore, the harvest rate information collected during these interviews was based on incomplete trips which documented only part of the total effort and harvest for these fishing trips (Robson 1961 \& 1991, Pollock et al. 1994). The use of roving survey methods introduced a sampling bias because the probability of interviewing a group is proportional to the duration of their fishing trip. That is, parties that fish for longer time periods are more likely to be encountered by field staff moving through the fishery, termed the 'length-of-stay' bias (Robson 1991, Pollock et al. 1994, Pollock et al. 1997, Hoenig et al. 1997), which means that harvest rates and discard rates derived from roving survey methods tend to be based on samples that contain an over-representative number of longer trips and an under-representative number of short trips. Roving survey methods require the following assumptions be made: (a) the harvest rate for the portion of fishing trip documented is the same as the harvest rate for the entire trip; and (b) the harvest rate of interviewed fishing parties is representative of the whole fishing population, which is the expected outcome for estimates derived from randomly selected samples (Malvestuto 1983, Phippen and Bergersen 1991, Pollock et al. 1994, Hoenig et al. 1997). When the objective is to estimate total harvest, and the interviews are based on incomplete trips, the correct harvest rate estimator to use is the 'mean of ratios' (Jones et al. 1995, Pollock et al. 1997, Hoenig et al. 1997). This estimator is essentially the mean of the individual harvest rates for all fishers interviewed on a given day. The 'mean of ratios' was used for estimating the harvest of the shore-based fishery. Hoenig et al. (1997) used simulation procedures to show that the 'mean of ratios' estimator has a large variance caused by the inclusion
of high harvest rates resulting from very short, incomplete trips that have harvested some fish already. These authors found that the truncation (exclusion) of all short incomplete trips reduced the variance greatly without inducing an appreciable bias. Hoenig et al. (1997) recommended the truncation of short trips less than 20-30 minutes but noted that there was a trade-off between the level of truncation used and the number of interviews that were discarded. We examined the relationship between the harvest rate and the duration of the fishing trip for shore-based interviews to determine the most appropriate level of truncation. We found that by discarding all incomplete trips that had been in progress for less than 30 fisher minutes, we were able to remove the interviews with the most extreme harvest rates and hence minimise the variance of the harvest rate estimator. The adoption of this truncation criterion resulted in the removal from harvest calculations of 104 (7.2\%) usable shore-based interviews from the first survey period and 171 (7.5\%) usable shore-based interviews from the second survey period. We had routinely asked shore-based fishing parties about the intended finishing time for their current trip. We retained and used shore-based interviews with fishing parties that had completed their trips but had fished for less than 30 fisher minutes. We believe it is logical to keep and use the data from these complete short trips, regardless of the small amount of time fished or the amount of harvest taken, because it is these short trips that are under-represented in roving surveys due to "length-of-stay" bias. The mean daily 'mean of ratios' estimator calculated for each base stratum was used for estimating the harvest of the shore-based fishery.

Seasonal harvest rates were calculated by combining estimates derived from day-type strata within each season. The contribution of each day-type stratum to the estimated seasonal harvest rate was weighted by the relative size of each day-type stratum within the season (Pollock et al. 1994). This means that a greater weighting was given to the weekday stratum because there are more weekdays in a month than there are weekend days in a month.

### 3.5.3. Harvest estimation for the boat-based and shore-based fisheries

The complemented survey designs used to assess the recreational fisheries used different on-site, contact methods to estimate effort and catch. Harvest estimation in the boat-based fishery used interviews derived from completed trips, whereas the shore-based fishery used interviews derived from incomplete trips. Thus, boat-based harvest was calculated as the product of boat-based effort and the mean daily 'ratio of means' harvest rate. Shore-based harvest was calculated as the product of shore-based effort and the truncated mean daily 'mean of ratios' harvest rate. Harvest estimation for both the boat-based and shore-based fisheries was done for each day-type stratum in a season. Seasonal estimates of harvest for the boat and shore fisheries were obtained by adding the estimates from the day-type strata together. Total fishery harvest estimates for each season were calculated by adding the boat-based and shore-based harvests together. Whenever strata were combined their variances were additive.

We did not attempt to make expanded estimates of harvest for any taxon that was considered to have been "rare" throughout the survey period - defined as any taxon that had been recorded from two or less interviews during a survey year, regardless of the number of individuals harvested in those trips. This definition of rarity was applied separately during each survey year to the boatbased and shore-based fisheries. All taxa which did not meet the criterion for rarity were classified as common taxa and expanded estimates of harvest were made for these taxa.

We converted the length measurements of fish, cephalopods and crabs taken during interviews into weights using length to weight keys (see Steffe and Chapman 2003). The remaining unmeasured component of the harvest (i.e. those fish seen during interviews but only counted) were assigned the median weight for that taxon as calculated from the pooled interview data for each season within a survey year. We used a median weight rather than a mean weight (as is traditionally done in angler surveys) because many of the estimated weight frequency distributions were highly
skewed, making the median a better estimate of the centre of the population (Sokal and Rohlf 1969). In some cases, the use of a mean would have resulted in higher estimates of harvest. We calculated medians separately for the boat-based and shore-based fisheries. When no measurements had been made for a taxon in a particular fishery (e.g. the boat fishery), we used the available measurements from the other fishery (e.g. the shore fishery). In some cases, measurements were not available for some taxa and so we could not estimate weights.

### 3.6. Statistical comparisons between survey periods

Annual estimates of recreational fishing effort and harvest (total fishery, boat-based fishery and shore-based fishery) and seasonal estimates of harvest rates for the boat-based and shore-based fisheries have been made for each survey period. We have presented $95 \%$ confidence limits for each of these estimated values. The $95 \%$ confidence limits provide information about the plausible range that contains the true value of the parameter that has been estimated. Thus, when comparing any two estimates of interest it is important to determine whether the confidence intervals overlap. When the confidence intervals overlap we cannot be $95 \%$ certain that the two estimates being compared are different. Thus, we conclude that in this case there is no statistically significant difference between the two estimates ( $\mathrm{p}>0.05$ ). Conversely, when the confidence intervals do not overlap we can be $95 \%$ certain that the two estimates are different. Thus, we can conclude that a statistically significant difference exists $(\mathrm{p}<0.05)$ between the two estimates.

### 3.7. Indicators of recreational fishing quality

An assessment of a recreational fishery can be improved if reliable indicators of fishing quality are available. We present two indicators of recreational fishing quality for the boat-based and shorebased fisheries in the Lake Macquarie fishery so that comparisons can be made between survey periods. The indicators are: (1) recreational harvest rates for the main species of recreational importance as determined by their relative harvest sizes in each survey year; and (2) size-frequency distributions for these same species. The harvest rates are based on calculations made using total fishing effort (non-directed effort) for a stratum. We present harvest rates for the boat-based and shore-based fisheries for each season and for each of the three Lake areas in units of number of fish per fisher hour. The amalgamation of these harvest rate data into larger groupings (e.g. annual or total Lake harvest rates) were not done for any taxon because they mask the trends seen at smaller spatial and temporal scales and do not enhance the assessment of the recreational fisheries. Size frequency distributions are presented for the entire fishery (boat and shore fisheries combined) during each of the two survey years.

## 4. RESULTS

### 4.1. $\quad$ Recreational fishing effort

We estimated that about 970,500 and 993,300 fisher hours of daytime recreational effort (boat and shore fisheries combined) was expended in the Lake Macquarie fishery during the first and second survey years respectively (Table 2). This represents an overall increase in recreational fishing effort of about $2.3 \%$ (no significant difference, $\mathrm{p}>0.05$ ) since the first survey (Table 2).

We estimated that about 681,800 and 769,300 fisher hours of daytime recreational boat-based effort was expended in the Lake Macquarie fishery during the first and second survey years respectively (Table 2). This represents an overall increase in boat-based recreational fishing effort of about $12.8 \%$ (significant difference, $\mathrm{p}<0.05$ ) since the first survey (Table 2). The boat-based effort accounted for $70.3 \%$ and $77.4 \%$ of the annual effort for the total fishery (boat and shore combined) during the first and second survey years respectively (Table 2).

We estimated that about 288,700 and 224,000 fisher hours of daytime recreational shore-based effort was expended in the Lake Macquarie fishery during the first and second survey years respectively (Table 2). This represents an overall decrease in shore-based recreational fishing effort of about $22.4 \%$ (significant difference, $\mathrm{p}<0.05$ ) since the first survey (Table 2). The shore-based effort accounted for $29.7 \%$ and $22.6 \%$ of the annual effort for the total fishery (boat and shore combined) during the first and second survey years respectively (Table 2).

### 4.2. Recreational harvest

### 4.2.1. Whole fishery

We recorded 60 taxa in the retained catch of recreational fishers (boat and shore combined) during the first survey year and 62 taxa during the second survey year (Table 3, Appendix 2). We estimated that about 543,700 fish, crabs and cephalopods ( 472,174 to 615,148 individuals approximate $95 \%$ Confidence Limits) were harvested by daytime recreational fishers from the Lake Macquarie fishery during the first survey year and about 497,500 fish, crabs and cephalopods ( 461,356 to 533,548 individuals - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 3). The crab and cephalopod component of the harvest was quite large, accounting for $36.1 \%$ of the harvest (about 196,200 individuals $-147,294$ to 245,192 approximate $95 \%$ Confidence Limits) during the first survey year and $19.8 \%$ of the harvest (about 98,600 individuals $-84,796$ to 112,426 approximate $95 \%$ Confidence Limits) during the second survey year (Table 3). The finfish component of the harvest accounted for $63.9 \%$ of the harvest (about 347,400 individuals $-295,317$ to 399,519 approximate $95 \%$ Confidence Limits) during the first survey year and $80.2 \%$ of the harvest (about 398,800 individuals - 365,493 to 432,189 approximate $95 \%$ Confidence Limits) during the second survey year (Table 3). In both survey years the recreational harvest was dominated by relatively few taxa (Table 3). The ten most commonly harvested taxa, by number, accounted for $88.0 \%$ and $93.2 \%$ of the daytime recreational harvest during the first and second survey years respectively (Table 3). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 3). For example, the total harvest (by number) of trumpeter whiting, dusky flathead, tailor and sand whiting, increased significantly since the first survey period (Table 3). In contrast, the total harvest (by number) of common squid, yellow-finned leatherjacket, and sand mullet decreased significantly since the first survey period (Table 3). Changes in total harvest (increases or
decreases) were also observed for luderick, blue swimmer crab, yellowfin bream, large-toothed flounder and flat-tail mullet but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 3).

We estimated that about 178 tonnes of fish, crabs and cephalopods (152 to 205 tonnes approximate $95 \%$ Confidence Limits) were harvested by daytime recreational fishers from the Lake Macquarie fishery during the first survey year and about 224 tonnes of fish, crabs and cephalopods (203 to 244 tonnes - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 4). The crab and cephalopod component of the harvest was quite large, accounting for $37.4 \%$ of the harvest (about 67 tonnes - 44 to 89 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $20.0 \%$ of the harvest (about 45 tonnes -37 to 53 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 4). The finfish component of the harvest accounted for $62.6 \%$ of the harvest (about 112 tonnes - 97 to 126 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $80.0 \%$ of the harvest (about 179 tonnes - 160 to 197 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 4). In both survey years the recreational harvest was dominated by relatively few taxa (Table 4). The ten most commonly harvested taxa, by weight, accounted for $90.1 \%$ and $93.5 \%$ of the daytime recreational harvest during the first and second survey years respectively (Table 4). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 4). For example, the total harvest (by weight) of dusky flathead, tailor, trumpeter whiting and sand whiting, increased significantly since the first survey period (Table 4). In contrast, the total harvest (by weight) of common squid, yellow-finned leatherjacket, and sand mullet was significantly less in the second survey period (Table 4). Changes in total harvest (increases or decreases) were also observed for luderick, blue swimmer crab, yellowfin bream, snapper and Australian salmon but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 4).

### 4.2.2. Boat fishery

We recorded 46 taxa in the retained catch of boat-based recreational fishers during the first survey year and 55 taxa during the second survey year (Table 5, Appendix 3). We estimated that about 369,100 fish, crabs and cephalopods (304,983 to 433,279 individuals - approximate $95 \%$ Confidence Limits) were harvested by daytime recreational fishers from the Lake Macquarie fishery during the first survey year and about 378,200 fish, crabs and cephalopods $(348,390$ to 407,972 individuals - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 5). The crab and cephalopod component of the harvest was quite large, accounting for $49.8 \%$ of the harvest (about 183,900 individuals $-135,321$ to 232,399 approximate $95 \%$ Confidence Limits) during the first survey year and $23.7 \%$ of the harvest (about 89,500 individuals $-76,735$ to 102,337 approximate $95 \%$ Confidence Limits) during the second survey year (Table 5). The finfish component of the harvest accounted for $50.2 \%$ of the harvest (about 185,300 individuals $-143,332$ to 227,210 approximate $95 \%$ Confidence Limits) during the first survey year and $76.3 \%$ of the harvest (about 288,600 individuals - 261,744 to 315,546 approximate $95 \%$ Confidence Limits) during the second survey year (Table 5). In both survey years the boat-based recreational harvest was dominated by relatively few taxa (Table 5). The ten most commonly harvested taxa, by number, accounted for $90.0 \%$ and $95.1 \%$ of the daytime recreational harvest during the first and second survey years respectively (Table 4). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 5). For example, the total boat-based harvest (by number) of trumpeter whiting, yellowfin bream, dusky flathead and tailor have increased significantly since the first survey period (Table 5). In contrast, the total boat-based harvest (by number) of common squid and yellowfinned leatherjacket has decreased significantly since the first survey period (Table 5). Changes in total harvest (increases or decreases) were also observed for blue swimmer crab, sand whiting,
luderick, large-toothed flounder, snapper and river garfish but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 5 ).
We estimated that about 116 tonnes of fish, crabs and cephalopods ( 92 to 141 tonnes - approximate $95 \%$ Confidence Limits) were harvested by daytime boat-based recreational fishers from the Lake Macquarie fishery during the first survey year and about 161 tonnes of fish, crabs and cephalopods (148 to 175 tonnes - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 6). The crab and cephalopod component of the harvest was quite large, accounting for $54.1 \%$ of the harvest (about 63 tonnes - 41 to 85 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $26.2 \%$ of the harvest (about 42 tonnes -35 to 50 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 6). The finfish component of the harvest accounted for $45.9 \%$ of the harvest (about 53 tonnes - 43 to 64 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $73.8 \%$ of the harvest (about 119 tonnes - 108 to 130 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 6). In both survey years the recreational boat-based harvest was dominated by relatively few taxa (Table 6). The ten most commonly harvested taxa, by weight, accounted for $92.7 \%$ and $93.8 \%$ of the daytime recreational harvest during the first and second survey years respectively (Table 6). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 6). For example, the total boat-based harvest (by weight) of dusky flathead, tailor, yellowfin bream, trumpeter whiting and sand whiting, have increased significantly since the first survey period (Table 6). In contrast, the total boat-based harvest (by weight) of common squid decreased significantly since the first survey period (Table 6). Changes in total harvest (increases or decreases) were also observed for blue swimmer crab, luderick, snapper, Australian salmon, yellow-finned leatherjacket and sand mullet but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 6 ).

### 4.2.3 $\quad$ Shore fishery

We recorded 38 taxa in the retained catch of shore-based recreational fishers during the first survey year and 44 taxa during the second survey year (Table 7, Appendix 4). We estimated that about 174,500 fish, crabs and cephalopods ( 142,978 to 206,082 individuals - approximate $95 \%$ Confidence Limits) were harvested by daytime recreational fishers from the Lake Macquarie fishery during the first survey year and about 119,300 fish, crabs and cephalopods ( 98,889 to 139,653 individuals - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 7). The crab and cephalopod component of the shore-based harvest was quite small, accounting for $7.1 \%$ of the harvest (about 12,400 individuals $-6,061$ to 18,705 approximate $95 \%$ Confidence Limits) during the first survey year and $7.6 \%$ of the harvest (about 9,100 individuals $-3,879$ to 14,271 approximate $95 \%$ Confidence Limits) during the second survey year (Table 7). The finfish component of the shore-based harvest accounted for $92.9 \%$ of the harvest (about 162,100 individuals - 131,235 to 193,059 approximate $95 \%$ Confidence Limits) during the first survey year and $92.4 \%$ of the harvest (about 110,200 individuals $-90,488$ to 129,904 approximate $95 \%$ Confidence Limits) during the second survey year (Table 7). In both survey years the shore-based recreational harvest was dominated by relatively few taxa (Table 7). The ten most commonly harvested taxa, by number, accounted for $91.2 \%$ and $92.6 \%$ of the daytime recreational harvest during the first and second survey years respectively (Table 7). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 7). For example, the total shore-based harvest (by number) of trumpeter whiting and dusky flathead have increased significantly since the first survey period (Table 7). In contrast, the total shore-based harvest (by number) of sand mullet, six-spined leatherjacket and fan-bellied leatherjacket has decreased significantly since the first survey period (Table 7). Changes in total harvest (increases or decreases) were also observed for luderick, yellowfin bream, common squid, tailor, flat-tail mullet, yellow-finned leatherjacket, sand whiting, and tarwhine but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 7 ).

We estimated that about 62 tonnes of fish, crabs and cephalopods ( 51 to 73 tonnes - approximate $95 \%$ Confidence Limits) were harvested by daytime shore-based recreational fishers from the Lake Macquarie fishery during the first survey year and about 62 tonnes of fish, crabs and cephalopods ( 47 to 77 tonnes - approximate $95 \%$ Confidence Limits) were harvested during the second survey year (Table 8). The crab and cephalopod component of the harvest was relatively small, accounting for $6.2 \%$ of the shore-based harvest (about 4 tonnes - 1 to 7 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $4.1 \%$ of the shore-based harvest (about 3 tonnes -1 to 4 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 8). The finfish component of the harvest accounted for $93.8 \%$ of the shore-based harvest (about 58 tonnes -48 to 69 tonnes approximate $95 \%$ Confidence Limits) during the first survey year and $95.9 \%$ of the harvest (about 60 tonnes - 45 to 75 tonnes approximate $95 \%$ Confidence Limits) during the second survey year (Table 8). In both survey years the recreational shore-based harvest was dominated by relatively few taxa (Table 8). The ten most commonly harvested taxa, by weight, accounted for $95.3 \%$ and $95.6 \%$ of the daytime recreational shore-based harvest during the first and second survey years respectively (Table 8). However, the composition and relative contribution of these dominant taxa changed markedly between survey years (Table 8). For example, the total shorebased harvest (by weight) of trumpeter whiting was significantly greater in the second survey period (Table 8). In contrast, the total shore-based harvest (by weight) of yellow-finned leatherjacket and sand mullet decreased significantly since the first survey period (Table 8). Changes in total harvest (increases or decreases) were also observed for luderick, yellowfin bream, dusky flathead, tailor, flat-tail mullet, common squid, blue swimmer crab, tarwhine, sand whiting, southern calamari and fan-bellied leatherjacket but these observed changes were not statistically different ( $\mathrm{p}>0.05$ ) between the survey periods (Table 8 ).
Table 2.
Estimates of daytime recreational fishing effort (fisher hours) with $95 \%$ confidence intervals for the boat-based, shore-based and total fisheries
for in Lake Macquarie for each survey year. The proportional changes between survey years and their statistical significance are presented.

| Fishery | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | Effort (fisher hours) | 95\% Confidence Intervals | \% total | Effort (fisher hours) | 95\% Confidence Intervals | \% total | \% Change (fisher hours) | Statistical Significance |
| Boat | 681,822 | 621,875 to 741,769 | 70.3\% | 769,251 | 730,293 to 808,209 | 77.4\% | 12.8\% | ns |
| Shore | 288,662 | 260,326 to 316,988 | 29.7\% | 224,029 | 205,156 to 242,902 | 22.6\% | -22.4\% | * |
| Total <br> Fishery | 970,484 | 904,181 to 1,036,787 | 100\% | 993,280 | $\mathbf{9 4 9 , 9 9 1}$ to $\mathbf{1 , 0 3 6 , 5 6 9}$ | 100\% | 2.3\% | ns |

* Significantly different, $\mathrm{p}<0.05$.
ns No significant difference, $\mathrm{p}>0.05$
Annual harvest estimates (number of individuals) with $95 \%$ confidence intervals for taxa taken by recreational fishers (boat-based and shorebased fishers combined) in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their statistical significance are presented.

| COMMON NAME | TOTAL HARVEST FOR WHOLE FISHERY |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |
|  | Total Fish (number) | $95 \%$ Confidence Intervals | \% Total | Total Fish (number) | 95\% Confidence Intervals | \% Total |
| Trumpeter Whiting | 63,413 | 27,709 to 99,117 | 11.7\% | 139,382 | 115,353 to 163,411 | 28.0\% |
| Luderick | 62,278 | 46,983 to 77,573 | 11.5\% | 62,492 | 44,336 to 80,648 | 12.6\% |
| Blue Swimmer Crab | 111,472 | 71,885 to 151,059 | 20.5\% | 61,451 | 50,455 to 72,447 | 12.4\% |
| Yellowfin Bream | 48,874 | 32,497 to 65,251 | 9.0\% | 56,792 | 49,903 to 63,681 | 11.4\% |
| Dusky Flathead | 18,886 | 13,191 to 24,581 | 3.5\% | 41,424 | 36,035 to 46,813 | 8.3\% |
| Tailor | 12,890 | 6,794 to 18,986 | 2.4\% | 39,415 | 31,093 to 47,737 | 7.9\% |
| Common Squid | 79,869 | 51,434 to 108,304 | 14.7\% | 34,724 | 26,447 to 43,001 | 7.0\% |
| Sand Whiting | 5,341 | 1,241 to 9,441 | 1.0\% | 13,806 | 9,883 to 17,729 | 2.8\% |
| Large-Toothed Flounder | 4,948 | 2,436 to 7,460 | 0.9\% | 7,400 | 5,971 to 8,829 | 1.5\% |
| Flat-Tail Mullet | 12,554 | 2,315 to 22,793 | 2.3\% | 6,392 | 2,456 to 10,328 | 1.3\% |
| Snapper | 6,499 | 4,264 to 8,734 | 1.2\% | 5,174 | 3,987 to 6,361 | 1.0\% |
| Yellow-Finned Leatherjacket | 32,475 | 15,500 to 49,450 | 6.0\% | 4,708 | 1,584 to 7,832 | 0.9\% |
| Small-Toothed Flounder | 3,693 | 2,237 to 5,149 | 0.7\% | 3,148 | 2,249 to 4,047 | 0.6\% |
| River Garfish | 6,393 | 0 to 14,238 | 1.2\% | 3,001 | 1,440 to 4,562 | 0.6\% |
| Sand Mullet | 35,063 | 17,921 to 52,205 | 6.4\% | 2,270 | 750 to 3,790 | 0.5\% |
| Tarwhine | 5,793 | 1,169 to 10,417 | 1.1\% | 2,085 | 1,240 to 2,930 | 0.4\% |
| Yellowtail | \#1 | - | <0.1\% | 1,677 | 177 to 3,177 | 0.3\% |
| Fan-Bellied Leatherjacket | 7,877 | 3,607 to 12,147 | 1.4\% | 1,556 | 251 to 2,861 | 0.3\% |
| Six-Spined Leatherjacket | 9826 | 4,645 to 15,007 | 1.8\% | 1,495 | 344 to 2,646 | 0.3\% |
| Southern Calamari | 2470 | 0 to 5,772 | 0.5\% | 1,264 | 299 to 2,229 | 0.3\% |
| Silver Trevally | 319 | 26 to 612 | <0.1\% | 1,118 | 431 to 1,805 | 0.2\% |
| Sea Garfish | 2,072 | 0 to 5,224 | 0.4\% | 840 | 0 to 1,915 | 0.2\% |

Table 3, continued.

| COMMON NAME | TOTAL HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEENSURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | Total Fish (number) | 95\% Confidence Intervals | \% Total | Total Fish (number) | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Octopus $\dagger$ | - | - | - | 755 | 85 to 1,425 | 0.2\% | - | - |
| Sea Mullet | \#2 | - | <0.1\% | 755 | 86 to 1,424 | 0.2\% | - | - |
| Southern Herring | 3,990 | 1,351 to 6,629 | 0.7\% | 701 | 0 to 1,409 | 0.1\% | -82.4\% | ns |
| Australian Salmon | \#1 | - | <0.1\% | 604 | 34 to 1,174 | 0.1\% | - | - |
| Chinaman Leatherjacket | - | - | - | 497 | 69 to 925 | <0.1\% | - | - |
| Stout Longtom $\dagger$ | 480 | 62 to 898 | <0.1\% | 455 | 117 to 793 | <0.1\% | -5.2\% | ns |
| Mud Crab | 2,430 | 0 to 5,492 | 0.4\% | 402 | 158 to 646 | <0.1\% | -83.5\% | ns |
| Slimy Mackerel | - | - | - | 282 | 0 to 696 | <0.1\% | - | - |
| Rough Leatherjacket | 300 | 0 to 666 | <0.1\% | 271 | 59 to 483 | <0.1\% | -9.7\% | ns |
| Silver Batfish $\dagger$ | \#1 | - | <0.1\% | 223 | 0 to 514 | <0.1\% | - | - |
| Striped Seapike | \#1 | - | $<0.1 \%$ | 218 | 0 to 444 | $<0.1 \%$ | - | - |
| Kingfish | \#3 | - | $<0.1 \%$ | 164 | 40 to 288 | $<0.1 \%$ | - | - |
| Mulloway | \#6 | - | <0.1\% | 149 | 0 to 321 | <0.1\% | - | - |
| School Whiting | 583 | 0 to 1,234 | 0.1\% | 141 | 0 to 304 | <0.1\% | -75.8\% | ns |
| Marbled Flathead | \#1 | - | <0.1\% | 130 | 0 to 286 | <0.1\% | - | - |
| Eastern Blue-Spotted Flathead | 688 | 11 to 1,365 | 0.1\% | \#2 | - | <0.1\% | - | - |
| Black Trevally (Spinefoot) $\dagger$ | 2,107 | 0 to 4,279 | 0.4\% | \#1 | - | <0.1\% | - | - |
| Other Taxa^ | \#62 | - | <0.1\% | \#88 | - | <0.1\% | - | - |
| Grand Total | 543,661 | 472,174 to 615,148 | 100.0\% | 497,452 | 461,356 to 533,548 | 100.0\% | -8.5\% | n.s. |
| Key: <br> \# Expanded estimates of har <br> Not recorded or not calcul <br> $\dagger$ Associated estimates of exp <br> $\wedge$ Other taxa details are prov <br> * Significantly different, $\mathrm{p}<0$ <br> ns No significant difference, p | not been cal re event occ ight (kg) are pendix 2. | ated. This observation rences. <br> provided for this taxo | classified <br> in Tables 6 | are event du <br> 8 because | g this time period and it <br> itable length to weight | occurrence <br> version key | mply noted. <br> not availab |  |

Annual harvest estimates (kilograms) with $95 \%$ confidence intervals for taxa taken by recreational fishers (boat-based and shore-based combined) in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their associated statistical significance are presented.

## Table 4.

| COMMON NAME | TOTAL HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  | $\begin{array}{r} \hline \text { Total Fish } \\ (\mathrm{kg}) \end{array}$ | 95\% Confidence Intervals | \% Total | $\begin{array}{r} \hline \text { Total Fish } \\ (\mathrm{kg}) \end{array}$ | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Luderick | 32,740 | 23,784 to 41,696 | 18.3\% | 42,768 | 28,083 to 57,453 | 19.1\% | 30.6\% | ns |
| Dusky Flathead | 12,088 | 8,760 to 15,416 | 6.8\% | 40,189 | 33,681 to 46,697 | 18.0\% | 232.5\% | * |
| Blue Swimmer Crab | 51,429 | 29,546 to 73,312 | 28.8\% | 37,707 | 30,053 to 45,361 | 16.9\% | -26.7\% | ns |
| Yellowfin Bream | 22,313 | 15,273 to 29,353 | 12.5\% | 32,994 | 28,761 to 37,227 | 14.8\% | 47.9\% | ns |
| Tailor | 4,082 | 2,116 to 6,048 | 2.3\% | 26,089 | 19,425 to 32,753 | 11.7\% | 539.2\% | * |
| Trumpeter Whiting | 5,985 | 2,887 to 9,083 | 3.4\% | 14,254 | 11,752 to 16,756 | 6.4\% | 138.2\% | * |
| Common Squid | 11,654 | 7,386 to 15,922 | 6.5\% | 5,653 | 4,254 to 7,052 | 2.5\% | -51.5\% | * |
| Sand Whiting | 1,344 | 448 to 2,240 | 0.8\% | 3,624 | 2,641 to 4,607 | 1.6\% | 169.7\% | * |
| Snapper | 4,341 | 1,265 to 7,417 | 2.4\% | 2,880 | 2,199 to 3,561 | 1.3\% | -33.6\% | ns |
| Australian Salmon | \#3 | - | $<0.1 \%$ | 2,641 | 0 to 5,937 | 1.2\% | - | - |
| Flat-Tail Mullet | 3,829 | 1,155 to 6,503 | 2.1\% | 2,609 | 1,044 to 4,174 | 1.2\% | -31.9\% | ns |
| Large-Toothed Flounder | 915 | 444 to 1,386 | 0.5\% | 1,793 | 1,444 to 2,142 | 0.8\% | 96.0\% | * |
| Silver Trevally | 184 | 0 to 384 | 0.1\% | 1,158 | 133 to 2,183 | 0.5\% | 528.6\% | ns |
| Mulloway | \#25 | - | <0.1\% | 1,129 | 0 to 2,417 | 0.5\% | - | - |
| Yellow-Finned Leatherjacket | 6,457 | 3,475 to 9,439 | 3.6\% | 1,116 | 427 to 1,805 | 0.5\% | -82.7\% | * |
| Southern Calamari | 2,052 | 0 to 5,081 | 1.2\% | 1,041 | 335 to 1,747 | 0.5\% | -49.3\% | ns |
| Tarwhine | 2,006 | 299 to 3,713 | 1.1\% | 962 | 525 to 1,399 | 0.4\% | -52.0\% | ns |
| Small-Toothed Flounder | 806 | 462 to 1,150 | 0.5\% | 835 | 581 to 1,089 | 0.4\% | 3.6\% | ns |
| Sand Mullet | 9,777 | 4,629 to 14,925 | 5.5\% | 629 | 145 to 1,113 | 0.3\% | -93.6\% | * |
| Sea Mullet | \#1 | - | <0.1\% | 587 | 0 to 1,304 | 0.3\% | - | - |
| Fan-Bellied Leatherjacket | 2,351 | 686 to 4,016 | 1.3\% | 566 | 248 to 884 | 0.3\% | -75.9\% | ns |
| Yellowtail | - | - | - | 468 | 0 to 953 | 0.2\% | - | - |

Table 4, continued.

| COMMON NAME | TOTAL HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | Total Fish (kg) | 95\% Confidence Intervals | \% Total | Total Fish (kg) | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Mud Crab | 1,583 | 0 to 3,401 | 0.9\% | 358 | 148 to 568 | 0.2\% | -77.4\% | ns |
| Kingfish | \#7 | - | <0.1\% | 340 | 88 to 592 | 0.2\% | - | - |
| Six-Spined Leatherjacket | 1,493 | 526 to 2,460 | 0.8\% | 324 | 109 to 539 | 0.1\% | -78.3\% | ns |
| River Garfish | 293 | 0 to 634 | 0.2\% | 180 | 72 to 288 | <0.1\% | -38.5\% | ns |
| Rough Leatherjacket | 104 | 0 to 245 | <0.1\% | 166 | 35 to 297 | <0.1\% | 59.3\% | ns |
| Chinaman Leatherjacket | - | - | - | 119 | 20 to 218 | <0.1\% | - | - |
| Marbled Flathead | \#1 | - | <0.1\% | 117 | 0 to 250 | <0.1\% | - | - |
| Sea Garfish | 107 | 0 to 276 | <0.1\% | 92 | 0 to 240 | $<0.1 \%$ | -13.6\% | ns |
| Striped Seapike | - | - | - | 78 | 0 to 167 | <0.1\% | - | - |
| Cobia | \#4 | - | <0.1\% | 60 | 60 to 60 | <0.1\% | - | - |
| Slimy Mackerel | - | - | - | 31 | 0 to 72 | <0.1\% | - | - |
| School Whiting | 83 | 0 to 176 | <0.1\% | 20 | 0 to 43 | <0.1\% | -75.9\% | ns |
| Southern Herring | 123 | 40 to 206 | <0.1\% | 8 | 0 to 16 | <0.1\% | -93.4\% | * |
| Eastern Blue-Spotted Flathead | 236 | 0 to 485 | 0.1\% | - | - | - | - | - |
| Other Taxa^ | \#12 | - | <0.1\% | \#9 | - | <0.1\% | - | - |
| Grand Total | 178,426 | 151,520 to 205,332 | 100.0\% | 223,594 | 203,413 to 243,775 | 100.0\% | 25.3\% | ns |

\#ey: Expanded estimates of harvest have not been calculated. This observation was classified as a rare event during this time period and its occurrence is simply noted. - Not recorded or not calculated for rare event occurrences.

- Other taxa details are provided in Appendix 2.
* $\quad$ Significantly different, $\mathrm{p}<0.05$.
ns

No significant difference, $\mathrm{p}>0.05$
Annual harvest estimates (number of individuals) with $95 \%$ confidence intervals for taxa taken by boat-based recreational fishers in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their associated statistical significance are presented.

| COMMON NAME | BOAT-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | Total Fish (number) | 95\% Confidence Intervals | \% Total | Total Fish (number) | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Trumpeter Whiting | 61,113 | 25,521 to 96,705 | 16.6\% | 129,580 | 105,802 to 153,358 | 34.3\% | 112.0\% | * |
| Blue Swimmer Crab | 110,069 | 70,514 to 149,624 | 29.8\% | 60,770 | 49,792 to 71,748 | 16.1\% | -44.8\% | ns |
| Yellowfin Bream | 24,560 | 16,296 to 32,824 | 6.7\% | 39,946 | 34,500 to 45,392 | 10.6\% | 62.6\% | * |
| Dusky Flathead | 16,934 | 11,363 to 22,505 | 4.6\% | 35,390 | 30,492 to 40,288 | 9.4\% | 109.0\% | * |
| Tailor | 8,675 | 4,375 to 12,975 | 2.4\% | 34,462 | 26,736 to 42,188 | 9.1\% | 297.3\% | * |
| Common Squid | 71,358 | 43,393 to 99,323 | 19.3\% | 27,151 | 20,616 to 33,686 | 7.2\% | -62.0\% | * |
| Sand Whiting | 4,963 | 884 to 9,042 | 1.3\% | 11,676 | 8,005 to 15,347 | 3.1\% | 135.3\% | ns |
| Luderick | 8,015 | 0 to 16,969 | 2.2\% | 8,180 | 4,363 to 11,997 | 2.2\% | 2.1\% | ns |
| Large-Toothed Flounder | 4,947 | 2,435 to 7,459 | 1.3\% | 6,960 | 5,579 to 8,341 | 1.8\% | 40.7\% | ns |
| Snapper | 6,233 | 4,035 to 8,431 | 1.7\% | 5,011 | 3,839 to 6,183 | 1.3\% | -19.6\% | ns |
| Small-Toothed Flounder | 3,692 | 2,236 to 5,148 | 1.0\% | 2,728 | 2,003 to 3,453 | 0.7\% | -26.1\% | ns |
| Flat-Tail Mullet | 3,884 | 786 to 6,982 | 1.1\% | 2,462 | 966 to 3,958 | 0.7\% | -36.6\% | ns |
| Yellow-Finned Leatherjacket | 11,568 | 5,225 to 17,911 | 3.1\% | 1,891 | 175 to 3,607 | 0.5\% | -83.7\% | * |
| Yellowtail | \#1 | - | $<0.1 \%$ | 1,674 | 174 to 3,174 | 0.4\% | - | - |
| River Garfish | 6,393 | 0 to 14,238 | 1.7\% | 1,376 | 355 to 2,397 | 0.4\% | -78.5\% | ns |
| Fan-Bellied Leatherjacket | 3,823 | 1,456 to 6,190 | 1.0\% | 1,361 | 84 to 2,638 | 0.4\% | -64.4\% | ns |
| Sea Garfish | 2,072 | 0 to 5,224 | 0.6\% | 830 | 0 to 1,905 | 0.2\% | -59.9\% | ns |
| Tarwhine | \#6 | - | $<0.1 \%$ | 804 | 316 to 1,292 | 0.2\% | - | - |
| Silver Trevally | - | - | - | 780 | 186 to 1,374 | 0.2\% | - | - |
| Octopus $\dagger$ | - | - | - | 628 | 0 to 1,282 | 0.2\% | - | - |
| Australian Salmon | - | - | - | 604 | 34 to 1,174 | 0.2\% | - | - |
| Southern Calamari | \#2 | - | $<0.1 \%$ | 576 | 182 to 970 | 0.2\% | - | - |

Table 5, continued.

| COMMON NAME | BOAT-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  |  | COMPARISON BETWEENSURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |  |
|  | Total Fish (number) | 95\% Confidence Intervals | \% Total | Total Fish (number) |  | $95 \%$ Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Chinaman Leatherjacket | - | - | - | 494 |  | 66 to 922 | 0.1\% | - | - |
| Mud Crab | 2,430 | 0 to 5,492 | 0.7\% | 399 |  | 155 to 643 | 0.1\% | -83.6\% | ns |
| Six-Spined Leatherjacket | 3,221 | 0 to 6,687 | 0.9\% | 383 |  | 100 to 666 | 0.1\% | -88.1\% | ns |
| Slimy Mackerel | - | - | - | 282 |  | 0 to 696 | <0.1\% | - | - |
| Sea Mullet | - | - | - | 272 |  | 0 to 742 | <0.1\% | - | - |
| Stout Longtom $\dagger$ | 226 | 0 to 509 | <0.1\% | 233 |  | 76 to 390 | <0.1\% | 3.1\% | ns |
| Sand Mullet | 13,333 | 2,188 to 24,478 | 3.6\% | 228 |  | 5 to 451 | <0.1\% | -98.3\% | * |
| Striped Seapike | \#1 | - | <0.1\% | 217 |  | 0 to 443 | <0.1\% | - | - |
| Rough Leatherjacket | 299 | 0 to 665 | $<0.1 \%$ | 195 |  | 27 to 363 | <0.1\% | - | - |
| Kingfish | \#2 | - | <0.1\% | 161 |  | 37 to 285 | <0.1\% | - | - |
| Mulloway | \#5 | - | <0.1\% | 149 |  | 0 to 321 | <0.1\% | - | - |
| School Whiting | 583 | 0 to 1,234 | 0.2\% | 141 |  | 0 to 304 | <0.1\% | -75.8\% | ns |
| Marbled Flathead | - | - | - | 130 |  | 0 to 286 | <0.1\% | - | - |
| Eastern Blue-Spotted Flathead | 688 | 11 to 1,365 | 0.2\% | \#2 |  | - | <0.1\% | - | - |
| Other Taxa^ | \#35 | - | <0.1\% | \#55 |  | - | $<0.1 \%$ | - | - |
| Grand Total | 369,131 | 304,983 to 433,279 | 100.0\% | 378,181 | 348,390 | 90 to 407,972 | 100.0\% | 2.5\% | ns |

[^0]Annual harvest estimates (kilograms) with $95 \%$ confidence intervals for taxa taken by boat-based recreational fishers in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their associated statistical significance are presented.
BOAT-BASED HARVEST FOR WHOLE FISHERY

| COMMON NAME | BOAT-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEENSURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  | $\begin{array}{r} \hline \text { Total Fish } \\ (\mathrm{kg}) \end{array}$ | 95\% Confidence Intervals | \% Total | $\begin{array}{r} \hline \text { Total Fish } \\ (\mathrm{kg}) \end{array}$ | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Blue Swimmer Crab | 50,913 | 29,038 to 72,788 | 43.8\% | 36,952 | 29,367 to 44,537 | 22.9\% | -27.4\% | ns |
| Dusky Flathead | 10,581 | 7,428 to 13,734 | 9.1\% | 36,508 | 30,239 to 42,777 | 22.6\% | 245.0\% | * |
| Tailor | 3,439 | 1,546 to 5,332 | 3.0\% | 24,384 | 17,852 to 30,916 | 15.1\% | 609.1\% | * |
| Yellowfin Bream | 11,406 | 6,996 to 15,816 | 9.8\% | 22,426 | 19,194 to 25,658 | 13.9\% | 96.6\% | * |
| Trumpeter Whiting | 5,730 | 2,650 to 8,810 | 4.9\% | 13,318 | 10,836 to 15,800 | 8.2\% | 132.4\% | * |
| Luderick | 4,623 | 0 to 10,512 | 4.0\% | 5,064 | 2,611 to 7,517 | 3.1\% | 9.5\% | ns |
| Common Squid | 10,354 | 6,184 to 14,524 | 8.9\% | 4,373 | 3,381 to 5,365 | 2.7\% | -57.8\% | * |
| Sand Whiting | 1,252 | 363 to 2,141 | 1.1\% | 3,079 | 2,165 to 3,993 | 1.9\% | 145.9\% | * |
| Snapper | 4,311 | 1,236 to 7,386 | 3.7\% | 2,840 | 2,161 to 3,519 | 1.8\% | -34.1\% | ns |
| Australian Salmon | - | - | - | 2,641 | 0 to 5,937 | 1.6\% | - | - |
| Large-Toothed Flounder | 915 | 444 to 1,386 | 0.8\% | 1,680 | 1,344 to 2,016 | 1.0\% | 83.7\% | ns |
| Mulloway | \#21 | - | - | 1,129 | 0 to 2,417 | 0.7\% | - | - |
| Flat-Tail Mullet | 1,664 | 212 to 3,116 | 1.4\% | 1,062 | 412 to 1,712 | 0.7\% | -36.2\% | ns |
| Small-Toothed Flounder | 806 | 462 to 1,150 | 0.7\% | 750 | 513 to 987 | 0.5\% | -7.0\% | ns |
| Silver Trevally | - | - | - | 697 | 0 to 1,601 | 0.4\% | - | - |
| Yellow-Finned Leatherjacket | 2,383 | 986 to 3,780 | 2.1\% | 602 | 46 to 1,158 | 0.4\% | -74.7\% | ns |
| Southern Calamari | \#2 | - | - | 555 | 130 to 980 | 0.3\% | - |  |
| Fan-Bellied Leatherjacket | 894 | 390 to 1,398 | 0.8\% | 526 | 213 to 839 | 0.3\% | -41.1\% | ns |
| Yellowtail | - | - | - | 468 | 0 to 953 | 0.3\% | - | - |
| Sea Mullet | - | - | - | 395 | 0 to 1,087 | 0.2\% | - | - |
| Mud Crab | 1,583 | 0 to 3,401 | 1.4\% | 355 | 145 to 565 | 0.2\% | -77.6\% | ns |
| Kingfish | \#2 | - | - | 330 | 78 to 582 | 0.2\% | - | - |

 29,367 to 44,537
 17,852 to 30,916
 10,836 to 15,800 2,611 to 7,517 $n$
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 | $\begin{array}{c}\text { SURVEY YEAR 1 } \\ \text { (March } \\ \text { 1999 to February 2000) }\end{array}$ |  |  |
| :---: | :---: | :---: |
| $\begin{array}{r}\text { Total Fish } \\ (\mathrm{kg})\end{array}$ | $\begin{array}{r}95 \% \text { Confidence } \\ \text { Intervals }\end{array}$ |  |$\quad \%$ Total $\begin{array}{rr}29,038 \text { to } 72,788 & 43.8 \% \\ 7,428 \text { to } 13,734 & 9.1 \%\end{array}$ 7,428 to 13,734

1,546 to 5,332
 0
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0
 444 to 1,386
 212 to 3,116
462 to 1,150 986 to 3,780

$\qquad$
 $1.4 \%$ ,

$$
6
$$

Table 6, continued.

| COMMON NAME | BOAT-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEEN <br> SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | Total Fish (kg) | 95\% Confidence Intervals | \% Total | Total Fish (kg) | 95\% Confidence Intervals | \% Total | \% Change <br> (number) | Statistical Significance |
| Tarwhine | \#1 | - | - | 294 | 119 to 469 | 0.2\% | - | - |
| Sand Mullet | 3,963 | 0 to 7,958 | 3.4\% | 163 | 0 to 335 | 0.1\% | -95.9\% | ns |
| Six-Spined Leatherjacket | 523 | 0 to 1,090 | 0.4\% | 151 | 42 to 260 | <0.1\% | -71.0\% | ns |
| Rough Leatherjacket | 104 | 0 to 245 | <0.1\% | 131 | 14 to 248 | <0.1\% | 25.9\% | ns |
| Chinaman Leatherjacket | - | - | - | 119 | 20 to 218 | <0.1\% | - | - |
| Marbled Flathead | - | - | - | 117 | 0 to 250 | <0.1\% | - | - |
| Sea Garfish | 107 | 0 to 276 | $<0.1 \%$ | 91 | 0 to 239 | $<0.1 \%$ | -14.6\% | ns |
| Striped Seapike | - | - | - | 77 | 0 to 166 | <0.1\% | - | - |
| River Garfish | 293 | 0 to 634 | 0.3\% | 73 | 19 to 127 | <0.1\% | -75.2\% | ns |
| Slimy Mackerel | - | - | - | 31 | 0 to 72 | <0.1\% | - | - |
| School Whiting | 83 | 0 to 176 | <0.1\% | 20 | 0 to 43 | $<0.1 \%$ | -75.9\% | ns |
| Eastern Blue-Spotted Flathead | 236 | 0 to 485 | 0.2\% | - | - | - | - | - |
| Other Taxa^ | \#12 | - | <0.1\% | \#66 | - | <0.1\% | - | - |
| Grand Total | 116,198 | $\mathbf{9 1 , 5 4 5}$ to 140,851 | 100.0\% | 161,467 | 148,055 to 174,879 | 100.0\% | 39.0\% | * |

Key:
\# Expanded estimates of harvest have not been calculated. This observation was classified as a rare event during this time period and its occurrence is simply noted. Not recorded or not calculated for rare event occurrences.
人 Other taxa details are provided in Appendix 3.

* Significantly different, $\mathrm{p}<0.05$.
ns
Annual harvest estimates (number of individuals) and $95 \%$ confidence intervals for taxa taken by shore-based recreational fishers in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their associated statistical significance are presented.

| COMMON NAME | SHORE-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | Total Fish (number) | $95 \%$ Confidence Intervals | \% Total | Total Fish (number) | 95\% Confidence | \% Total | \% Change (number) | Statistical Significance |
| Luderick | 54,263 | 41,863 to 66,663 | 31.1\% | 54,312 | 36,562 to 72,062 | 45.5\% | 0.1\% | ns |
| Yellowfin Bream | 24,314 | 10,175 to 38,453 | 13.9\% | 16,846 | 12,628 to 21,064 | 14.1\% | -30.7\% | ns |
| Trumpeter Whiting | 2,300 | 0 to 5,123 | 1.3\% | 9,802 | 6,334 to 13,270 | 8.2\% | 326.2\% | * |
| Common Squid | 8,511 | 3,358 to 13,664 | 4.9\% | 7,573 | 2,493 to 12,653 | 6.3\% | -11.0\% | ns |
| Dusky Flathead | 1,952 | 765 to 3,139 | 1.1\% | 6,034 | 3,786 to 8,282 | 5.1\% | 209.1\% | * |
| Tailor | 4,215 | 0 to 8,536 | 2.4\% | 4,953 | 1,861 to 8,045 | 4.2\% | 17.5\% | ns |
| Flat-Tail Mullet | 8,670 | 0 to 18,429 | 5.0\% | 3,930 | 289 to 7,571 | 3.3\% | -54.7\% | ns |
| Yellow-Finned Leatherjacket | 20,907 | 5,162 to 36,652 | 12.0\% | 2,817 | 206 to 5,428 | 2.4\% | -86.5\% | ns |
| Sand Whiting | 378 | 0 to 797 | 0.2\% | 2,130 | 745 to 3,515 | 1.8\% | 463.5\% | ns |
| Sand Mullet | 21,730 | 8,705 to 34,755 | 12.5\% | 2,042 | 538 to 3,546 | 1.7\% | -90.6\% | * |
| River Garfish | - | - | - | 1,625 | 444 to 2,806 | 1.4\% | - | - |
| Tarwhine | 5,787 | 1,163 to 10,411 | 3.3\% | 1,281 | 591 to 1,971 | 1.1\% | -77.9\% | ns |
| Six-Spined Leatherjacket | 6,605 | 2,754 to 10,456 | 3.8\% | 1,112 | 0 to 2,228 | 0.9\% | -83.2\% | * |
| Southern Herring | 3,990 | 1,351 to 6,629 | 2.3\% | 701 | 0 to 1,409 | 0.6\% | -82.4\% | ns |
| Southern Calamari | 2,468 | 0 to 5,770 | 1.4\% | 688 | 0 to 1,569 | 0.6\% | -72.1\% | ns |
| Blue Swimmer Crab | 1,403 | 0 to 2,989 | 0.8\% | 681 | 52 to 1,310 | 0.6\% | -51.5\% | ns |
| Sea Mullet | \#2 | - | <0.1\% | 483 | 7 to 959 | 0.4\% | - | - |
| Large-Toothed Flounder | \#1 | - | <0.1\% | 440 | 75 to 805 | 0.4\% | - | - |
| Small-Toothed Flounder | \#1 | - | $<0.1 \%$ | 420 | 0 to 951 | 0.4\% | - | - |
| Silver Trevally | 319 | 26 to 612 | 0.2\% | 338 | 0 to 684 | 0.3\% | 6.0\% | ns |
| Silver Batfish $\dagger$ | \#1 | - | <0.1\% | 223 | 0 to 514 | 0.2\% | - | - |
| Stout Longtom $\dagger$ | 254 | 0 to 563 | 0.1\% | 222 | 0 to 522 | 0.2\% | -12.6\% | ns |

Table 7, continued.

| COMMON NAME | SHORE-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  |  | COMPARISON BETWEENSURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |  |
|  | Total Fish (number) | 95\% Confidence | \% Total | Total Fish (number) |  | $95 \%$ Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Fan-Bellied Leatherjacket | 4054 | 500 to 7,608 | 2.3\% | 195 |  | 0 to 463 | 0.2\% | -95.2\% | * |
| Snapper | 266 | 0 to 674 | 0.2\% | 163 |  | 0 to 352 | 0.1\% | -38.7\% | ns |
| Octopus $\dagger$ | - | - | - | 127 |  | 0 to 275 | 0.1\% | - | - |
| Rough Leatherjacket | \#1 | - | <0.1\% | 76 |  | 0 to 205 | <0.1\% | - | - |
| Black Trevally (Spinefoot) $\dagger$ | 2,107 | 0 to 4,279 | 1.2\% | \#1 |  | - | <0.1\% | - | - |
| Other Taxa^ | \#31 | - | <0.1\% | \#56 |  | - | $<0.1 \%$ | - | - |
| Grand Total | 174,530 | $\mathbf{1 4 2 , 9 7 8}$ to 206,082 | 100.0\% | 119,271 | 98,889 | 89 to 139,653 | 100.0\% | -31.7\% | ns |

Not recorded or not calculated for rare event occurrences.
Associated estimates of expanded weight $(\mathrm{kg})$ are not provided
Other taxa details are provided in Appendix 4.
ns No significant difference, $\mathrm{p}>0.05$
Annual harvest estimates (kilograms) with $95 \%$ confidence intervals for taxa taken by shore-based recreational fishers in the Lake Macquarie fishery for each survey year. The proportional changes between survey years and their associated statistical significance are presented.

| SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |
| ---: | ---: | ---: |
| Total Fish $95 \%$ Confidence <br> $(\mathrm{kg})$  | Intervals | \% Total |
|  |  |  |
| 37,704 | 23,225 to 52,183 | $60.7 \%$ |
| 10,568 | 7,834 to 13,302 | $17.0 \%$ |
| 3,682 | 1,932 to 5,432 | $5.9 \%$ |
| 1,705 | 388 to 3,022 | $2.7 \%$ |
| 1,548 | 124 to 2,972 | $2.5 \%$ |
| 1,280 | 293 to 2,267 | $2.1 \%$ |
| 937 | 622 to 1,252 | $1.5 \%$ |
| 755 | 0 to 1,782 | $1.2 \%$ |
| 668 | 267 to 1,069 | $1.1 \%$ |
| 545 | 185 to 905 | $0.9 \%$ |
| 514 | 108 to 920 | $0.8 \%$ |
| 486 | 0 to 1,049 | $0.8 \%$ |
| 466 | 13 to 919 | $0.7 \%$ |
| 461 | 0 to 944 | $0.7 \%$ |
| 192 | 5 to 379 | $0.3 \%$ |
| 173 | 0 to 358 | $0.3 \%$ |
| 113 | 19 to 207 | $0.2 \%$ |
| 108 | 15 to 201 | $0.2 \%$ |
| 85 | 0 to 177 | $0.1 \%$ |
| 41 | 0 to 90 | $<0.1 \%$ |
| 39 | 0 to 95 | $<0.1 \%$ |
| 35 | 0 to 95 | $<0.1 \%$ |
|  |  |  |


|  |
| ---: |
| \% Total |
| $45.2 \%$ |
| $17.5 \%$ |
| $2.4 \%$ |
| $1.0 \%$ |
| $3.5 \%$ |
| $2.1 \%$ |
| $0.4 \%$ |
| $0.8 \%$ |
| $3.2 \%$ |
| $0.1 \%$ |
| $6.5 \%$ |
| $3.3 \%$ |
| $9.3 \%$ |
| $0.3 \%$ |
| $<0.1 \%$ |
| $1.6 \%$ |
| - |
| - |
| - |
| $<0.1 \%$ |
| $2.3 \%$ |

 $95 \%$ Confidence
Intervals
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| SHORE-BASED HARVEST FOR WHOLE FISHERY |  |  |  | COMPARISON BETWEEN SURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: |
| SURVEY YEAR 1 (March 1999 to February 2000) | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
| Total Fish$(\mathrm{kg})$$\quad$$95 \%$ Confidence <br> Intervals$\quad \%$ Total | Total Fish (kg) | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance | $\begin{array}{cc} \\ 34.1 \% & \mathrm{~ns}\end{array}$


Table 8, continued.

| COMMON NAME | SHORE-BASED HARVEST FOR WHOLE FISHERY |  |  |  |  |  | COMPARISON BETWEENSURVEY YEARS |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  | SURVEY YEAR 1(March 1999 to February 2000) |  |  | SURVEY YEAR 2(December 2003 to November 2004) |  |  |  |  |
|  | $\begin{array}{r} \text { Total Fish } \\ (\mathrm{kg}) \end{array}$ | 95\% Confidence Intervals | \% Total | Total Fish $(\mathrm{kg})$ | 95\% Confidence Intervals | \% Total | \% Change (number) | Statistical Significance |
| Southern Herring | 123 | 40 to 206 | 0.2\% | 8 | 0 to 16 | $<0.1 \%$ | -93.4\% | * |
| Other Taxa^ | \#17 | - | <0.1\% | \#17 | - | $<0.1 \%$ | - | - |
| Grand Total | 62,228 | 51,451 to 73,005 | 100.0\% | 62,128 | 47,049 to 77,207 | 100.0\% | -0.2\% | ns |

\# Expanded estimates of harvest have not been calculated. This observation was classified as a rare event during this time period and its occurrence is simply noted.

* $\quad$ Significantly different, $\mathrm{p}<0.05$.
ns No significant difference, $\mathrm{p}>0.05$
- Not recorded or not calculated for rare event occurrences.


### 4.3. Indicators of Recreational Fishing Quality

### 4.3.1. $\quad$ Recreational harvest rates

Seasonal trends are evident in the harvest rate information, however, these data are highly variable which means that estimates of seasonal harvest rates are usually imprecise. Thus, most comparisons of harvest rates among seasons within a survey year or between survey years are not statistically significantly different (see Figs 2 to 23). A brief description of the harvest rate data that focuses on statistically detectable differences between survey periods is provided below for the main species of recreational importance.

### 4.3.1.1. $\quad$ Trumpeter whiting

In the Southern Lake area, the boat-based harvest rate observed during the Summer of the second survey year was significantly greater ( $\mathrm{p}<0.05$ ) than that measured during the Summer of the first survey year (Fig. 2). In the Southern Lake area, the shore-based harvest rates observed during all seasons of the second survey were found to be significantly higher than those measured during the corresponding seasons during the first survey year (Fig. 3).

### 4.3.1.2. Luderick

There were no statistically significant differences in seasonal harvest rates between survey periods for boat-based fishers (Fig. 4). In the Southern Lake area, the Spring harvest rate of shore-based fishers in the first survey year was significantly higher ( $\mathrm{p}<0.05$ ) than the corresponding season during the second survey year (Fig. 5).

### 4.3.1.3. Blue swimmer crab

There were no statistically significant differences in seasonal harvest rates between survey periods for boat-based fishers or shore-based fishers (Figs. 6 \& 7).

### 4.3.1.4. Yellowfin bream

In the Swansea Channel and the Southern Lake area, the boat-based harvest rates observed during Autumn of the second survey year were significantly greater ( $\mathrm{p}<0.05$ ) than those measured during the corresponding season in the same areas of the Lake during the first survey year (Fig. 8). In the Southern Lake area, the Winter harvest rate of boat-based fishers was significantly higher during the second survey year (Fig. 8). Also, in the Southern Lake area, the Summer harvest rate of shorebased fishers was significantly higher during the second survey year (Fig. 9).

### 4.3.1.5. Dusky flathead

In the Southern Lake area, the boat-based and shore-based harvest rates observed during Winter and Spring of the second survey year were significantly greater ( $\mathrm{p}<0.05$ ) than those measured during the corresponding seasons during the first survey year (Figs. $10 \& 11$ ). Also, the Summer harvest rate of boat-based fishers was significantly higher in the Southern Lake area during the second survey year (Fig. 10).

### 4.3.1.6. Tailor

In the Northern Lake area, the boat-based harvest rate observed during the Autumn of the second survey year was significantly greater $(\mathrm{p}<0.05)$ than that measured during the Autumn of the first survey year (Fig. 12). In the Southern Lake area, the boat-based harvest rates observed during the Autumn, Winter and Spring seasons of the second survey year were found to be significantly higher than those measured during the corresponding seasons during the first survey year (Fig. 12). There were no statistically significant differences in seasonal harvest rates between survey periods for shore-based fishers (Fig. 13).

### 4.3.1.7. Common squid

In the Southern Lake area, the boat-based harvest rates observed during Autumn and Winter of the first survey year were significantly greater $(\mathrm{p}<0.05)$ than those measured during the corresponding seasons of the second survey year (Fig. 14). For the shore-based fishery, harvest rates were significantly greater $(\mathrm{p}<0.05)$ in the Southern Lake area for Summer in the first survey year (Fig. 15).

### 4.3.1.8. $\quad$ Sand whiting

In the Northern Lake area, the boat-based harvest rates observed during Summer of the second survey year were significantly greater $(\mathrm{p}<0.05)$ than those measured during the Summer of the first survey year (Fig. 16). There were no statistically significant differences in seasonal harvest rates between survey periods for shore-based fishers (Fig.17).

### 4.3.1.9. Large-toothed flounder

In the Swansea Channel, the boat-based harvest rates observed during Summer of the second survey year were significantly greater ( $\mathrm{p}<0.05$ ) than those measured during the Summer of the first survey year (Fig. 18). There were no statistically significant differences in seasonal harvest rates between survey periods for shore-based fishers (Fig 19).

### 4.3.1.10. Yellow-finned leatherjacket

In the Northern Lake area, the boat-based harvest rates observed during Summer of the first survey year were significantly greater ( $\mathrm{p}<0.05$ ) than those measured during the Summer of the second survey year (Fig. 20). For the shore-based fishery, harvest rates were significantly greater ( $\mathrm{p}<0.05$ ) in the Northern Lake area for Autumn, Winter and Spring in the first survey year (Fig. 21).

### 4.3.1.11. Sand mullet

There were no statistically significant differences in seasonal harvest rates between survey periods for boat-based fishers or shore-based fishers (Figs. $22 \& 23$ ).


Figure 2. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for trumpeter whiting taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake for each survey year.


Figure 3. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for trumpeter whiting taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## LUDERICK - BOAT FISHERY

NORTHERN LAKE


Figure 4. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for luderick taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## LUDERICK - SHORE FISHERY

NORTHERN LAKE


Figure 5. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for luderick taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 6. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for blue swimmer crab taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 7. Recreational harvest rate estimates (fish per fisher hour) and $95 \%$ confidence intervals for blue-swimmer crab taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 8. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for yellowfin bream taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

YELLOWFIN BREAM - SHORE FISHERY


Figure 9. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for yellowfin bream taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 10. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for dusky flathead taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

DUSKY FLATHEAD - SHORE FISHERY


Figure 11. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for dusky flathead taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 12. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for tailor taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

TAILOR - SHORE FISHERY
NORTHERN LAKE


Figure 13. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for tailor taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## COMMON SQUID - BOAT FISHERY



Figure 14. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for common squid taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 15. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for common squid taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 16. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for sand whiting taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.


Figure 17. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for sand whiting taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## LARGE-TOOTHED FLOUNDER - BOAT FISHERY



Figure 18. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for large-toothed flounder taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## LARGE-TOOTHED FLOUNDER - SHORE FISHERY



Figure 19. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for large-toothed flounder taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## YELLOW-FINNED LEATHERJACKET

- BOAT FISHERY

NORTHERN LAKE


Figure 20. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for yellow-finned leatherjacket taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

## YELLOW-FINNED LEATHERJACKET - SHORE FISHERY



Figure 21. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for yellow-finned leatherjacket taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

SAND MULLET - BOAT FISHERY


Figure 22. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for sand mullet taken by boat-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

SAND MULLET - SHORE FISHERY


Figure 23. Recreational harvest rate estimates (fish per fisher hour) with $95 \%$ confidence intervals for sand mullet taken by shore-based fishers in the Northern Lake, Swansea Channel and Southern Lake areas for each survey year.

### 4.3.2. Size-frequency distributions

Descriptive statistics of all measurements taken for each taxon by boat-based and shore-based fishers during each survey period are presented in Appendices 2-4. Here, we present length frequency distributions and comparisons between survey periods (shore and boat fisheries combined) for the main species of recreational importance.

### 4.3.2.1. Trumpeter whiting

A comparison of the length frequency distributions between the two survey periods shows great similarity between survey years (Fig. 24). There was no change in the mean and median fork lengths of trumpeter whiting between survey years (Fig. 24).

### 4.3.2.2. Luderick

A comparison of the length frequency distributions between the two survey periods shows that luderick taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 25). Luderick harvested during the second survey year had larger mean and median fork lengths (Fig. 25).

### 4.3.2.3. Blue swimmer crab

A comparison of the length frequency distributions between the two survey periods shows that the crabs taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 26). Blue swimmer crabs harvested during the second survey year had larger mean and median carapace lengths (Fig. 26).

### 4.3.2.4. Yellowfin bream

A comparison of the length frequency distributions between the two survey periods shows that yellowfin bream taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 27). Yellowfin bream harvested during the second survey year had larger mean and median fork lengths (Fig. 27).

### 4.3.2.5. Dusky flathead

A comparison of the length frequency distributions between the two survey periods shows that dusky flathead taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 28). Dusky flathead harvested during the second survey year had larger mean and median fork lengths but it should be noted that the minimum legal length for this species was increased from 33 cm TL to 36 cm TL in the period between the surveys (Fig. 28).

### 4.3.2.6. Tailor

A comparison of the length frequency distributions between the two survey periods shows that tailor taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 29). Tailor harvested during the second survey year had larger mean and median fork lengths (Fig. 29).

### 4.3.2.7. Common Squid

A comparison of the length frequency distributions between the two survey periods shows that common squid taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 30). Squid harvested during the second survey year had larger mean and median mantle lengths (Fig. 30).

### 4.3.2.8. $\quad$ Sand whiting

A comparison of the length frequency distributions between the two survey periods shows that sand whiting taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 31). Sand whiting harvested during the second survey year had larger mean and median fork lengths (Fig. 31).

### 4.3.2.9. Large-toothed flounder

A comparison of the length frequency distributions between the two survey periods shows that large-toothed flounder taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 32). Large-toothed flounder harvested during the second survey year had larger mean and median total lengths (Fig. 32).

### 4.3.2.10. Yellow-finned leatherjacket

A comparison of the length frequency distributions between the two survey periods showed little similarity (Fig. 33). Yellow-finned leatherjackets harvested during the second survey year had the same mean length but smaller median total length indicating smaller fish were taken during the second survey year (Fig. 33).

### 4.3.2.11. Sand mullet

A comparison of the length frequency distributions between the two survey periods shows that sand mullet taken during the second survey year were, on average, larger than those harvested during the first survey year (Fig. 34). Sand mullet harvested during the second survey year had larger mean and median fork lengths (Fig. 34).

## TRUMPETER WHITING

Whole fishery (Boat and Shore combined)

Figure 24. Trumpeter whiting - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.



FORK LENGTH (cm)

LUDERICK
Whole fishery (Boat and Shore combined)

FORK LENGTH (cm)
Figure 25. Luderick - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.
BLUE SWIMMER CRAB
Whole fishery (Boat and Shore combined)

CARAPACE LENGTH (cm)
Figure 26. Blue swimmer crab - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.
 (March 1999 to February 2000)

CARAPACE LENGTH (cm)
CARAPACE LENGTH (cm)
COMPARISON BETWEEN

乙 0
®
\%
YELLOWFIN BREAM
Whole fishery (Boat and Shore combined)

FORK LENGTH (cm)

FORK LENGTH (cm)

[^1]DUSKY FLATHEAD
Whole fishery (Boat and Shore co

SURVEY YEAR 1

Figure 28. Dusky flathead - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake

## TAILOR <br> Whole fishery (Boat and Shore combined)


FORK LENGTH (cm)
Figure 29. Tailor - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.

FORK LENGTH (cm)

## COMMON SQUID <br> Whole fishery (Boat and Shore combined)



MANTLE LENGTH (cm)

## SAND WHITING <br> Whole fishery (Boat and Shore combined)


FORK LENGTH (cm)
Figure 31. Sand whiting - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake
LARGE-TOOTHED FLOUNDER


TOTAL LENGTH (cm)
Figure 32. Large-toothed flounder - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.
YELLOW-FINNED LEATHERJACKET Whole fishery (Boat and Shore combined)


TOTAL LENGTH (cm)
Figure 33. Yellow-finned leatherjacket - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake Macquarie.
TOTAL LENGTH (cm)



TOTAL LENGTH (cm)

SAND MULLET
Whole fishery (Boat and Shore combined)




## FORK LENGTH (cm)

Figure 34. Sand mullet - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake

## 5. DISCUSSION

The establishment of Lake Macquarie as a Recreational Fishing Haven (RFH) has changed the allocation of fisheries resources in this waterway between the recreational and commercial sectors. The removal of access for commercial fishers to Lake Macquarie occurred during May 2002 thereby creating additional recreational fishing opportunities. This report focuses on comparisons made between two separate daytime recreational fishing surveys of Lake Macquarie. The first annual survey was done during the pre-RFH period (March 1999 to February 2000) and the second annual survey was done during the post-RFH period (December 2003 to November 2004). These annual surveys provide a snapshot of the recreational fishery of Lake Macquarie before RFH implementation and after RFH implementation. However, the representativeness of these two unreplicated survey periods as measures of pre-RFH and post-RFH conditions within the recreational fishery of Lake Macquarie remains unknown.

The assessment of environmental disturbance or impacts arising from management interventions is made difficult because it is often uncertain whether a causal relationship exists between the management event (e.g. establishment of a RFH) that has occurred and any changes in fish populations or the recreational fishery that are measured at a later time. The changes in the recreational fishery that have been detected following the implementation of the RFH may be in part attributable to the impact of the management intervention and/or may be in part attributable to natural fluctuations in fish abundance and catchability. These can be large in an open system that allows migratory fish stocks to enter and leave the Lake. Nonetheless, the comparison between the two annual survey periods does show that real differences have occurred in the Lake Macquarie fishery since the first pre-RFH survey period.

Have there been changes in the recreational fishery since the exclusion of commercial fishing by the establishment of Lake Macquarie as a RFH? In an extractive fishery the estimation of harvest provides a direct measure of the impact of fishing. Thus, changes in the harvest (number and weight of fish, crabs and cephalopods) and the relative composition of the harvest between annual survey periods are important measures that were used to assess change in the recreational fishery through time. We found no significant difference in the total annual harvest of fish, crabs and cephalopods, by number or weight, between survey periods for the whole fishery (Tables $3 \& 4$ ). The recreational harvest in both survey years was dominated by relatively few taxa, however, the composition and relative contribution of these dominant taxa changed markedly between survey years (Tables $3 \& 4$ ). The recreational harvest of trumpeter whiting ( $119.8 \%$ by number, $138.2 \%$ by weight), dusky flathead ( $119.3 \%$ by number, $232.5 \%$ by weight), tailor ( $205.8 \%$ by number, $539.2 \%$ by weight) and sand whiting ( $158.5 \%$ by number, $169.7 \%$ by weight) was significantly greater ( $\mathrm{p}<0.05$ ) in the second survey year (Tables $3 \& 4$ ). In contrast, the recreational harvest of common squid ( $-56.5 \%$ by number, $-51.5 \%$ by weight), yellow-finned leatherjacket ( $-85.5 \%$ by number, $-82.7 \%$ by weight) and sand mullet ( $-93.5 \%$ by number, $-93.6 \%$ by weight) was significantly less ( $\mathrm{p}<0.05$ ) in the second survey year (Tables $3 \& 4$ ). Increased harvest levels during the second survey year were recorded for yellowfin bream ( $16.2 \%$ by number, $47.9 \%$ by weight) and luderick ( $0.3 \%$ by number, $30.6 \%$ by weight) but these changes were not statistically significant (Tables $3 \& 4$ ). Blue swimmer crab ( $-44.9 \%$ by number, $-26.7 \%$ by weight) and flat-tail mullet ( $-49.1 \%$ by number, $-31.9 \%$ by weight) had lower harvest levels during the second survey period but these changes were not statistically significant (Tables $3 \& 4$ ). The harvest of largetoothed flounder was greater in the second survey period ( $49.6 \%$ by number but not statistically significant, $96.0 \%$ by weight $\mathrm{p}<0.05$ ). These findings indicate that the post-RFH recreational fishery in Lake Macquarie was very different to the fishery that had existed prior to the implementation of the RFH.

A better understanding of these changes between survey periods in the Lake Macquarie recreational fishery can be achieved by considering the factors that influence the size of harvest levels and how they may have changed since the first survey year. The major factors that influence the size of the recreational harvest are fishing effort, harvest rates and the size of fish, crabs and cephalopods taken. A discussion of each of these main factors follows.

Fishing effort can influence the total harvest in two ways. Fishing effort can have a direct effect as measured by absolute changes in the time spent fishing (assuming harvest rate remains constant) and also an indirect effect which could be due to changes in the direction or targeting of fishing effort. The total fishing effort (boat and shore combined) in the fishery showed little change between survey periods (about $2.3 \%$ overall - not statistically significant - Table 2). However, different trends were evident in the boat-based and shore-based fisheries. The fishing effort expended in the larger boat-based fishery increased by about $12.8 \%$ during the second survey year but this change was not statistically significant. This additional boat-based fishing effort may have contributed to increases in the harvest of species that are targeted by boat-based drift fishing, such as, dusky flathead, large-toothed flounder and trumpeter whiting. Interestingly, there was a statistically significant reduction of about $22.4 \%$ in the level of fishing effort expended in the smaller shore-based fishery during the second survey year. This decreased fishing effort was mainly due to the exclusion of recreational fishers from the hot water outlet of the Eraring power station. This part of the Lake had been very popular with shore-based fishers during the Winter and Spring seasons of the first survey year. The hot water outlet site had been characterised previously by many large catches of luderick. The observed changes (increases or decreases) in harvest levels for different species cannot be explained by changes in effort alone. The proportional changes in recreational harvest between survey years are much larger than the corresponding proportional changes in fishing effort. This is true for both the boat-based and shore-based fisheries in Lake Macquarie.

Changes in targeting may also help explain changes in harvest between survey years. It is plausible that less favoured species are targeted by recreational fishers whenever it becomes difficult to catch their favoured species. This behaviour leads to the targeting of whatever is available at the time and usually occurs when favoured species are less accessible to the recreational fishery. For example, changes in targeting behaviour would be expected to shift away from favoured species during periods of low abundance, low catchability or when the available resource is being used heavily by many commercial and recreational users as in the case of the pre-RFH fishery in Lake Macquarie. Conversely, changes in targeting behaviour would be expected to shift towards favoured species during periods of high abundance, high catchability or when the fishing pressure on the available resource is reduced by excluding a large user-group (i.e. the commercial sector) as in the case of the post-RFH fishery in Lake Macquarie.

Do the seasonal harvest rate data for the boat-based and shore-based fisheries indicate any major changes in fishing quality since the first survey year? Seasonal trends are evident in the harvest rate data, however, these data are highly variable making it difficult to detect statistically significant differences between survey years (see Figs. 2 to 23). Trumpeter whiting, dusky flathead, tailor and sand whiting all had significant increases in recreational harvest, by number, during the second survey year. For these species, these observations were supported by the seasonal harvest rate comparisons made between the survey years. All of the significant harvest rate differences detected between corresponding seasons indicated that for these species the harvest rates were better during the post-RFH survey year. Common squid, yellow-finned leatherjacket and sand mullet all had significant decreases in recreational harvest, by number, during the second survey year. Seasonal harvest rate comparisons that were significantly different for common squid and yellow-finned leatherjacket indicated lower harvest rates during the post-RFH year, with the exception of a comparison between Autumn seasons which indicated that the harvest rate of common squid in the Swansea Channel area was better during the second year. The harvest rate data for sand mullet
were too imprecise for the detection of any significant seasonal differences. Yellowfin bream and large-toothed flounder harvests were estimated to be greater during the second year but these trends were not statistically significantly different between years. Even so, all of the significant harvest rate differences detected between corresponding seasons indicated that for these two species, the harvest rates were better during the post-RFH survey year. The estimated harvest of blue swimmer crab had decreased since the first survey year but this trend was not statistically significant. Unfortunately, the harvest rate data for blue swimmer crab were too imprecise for the detection of any significant seasonal differences. The recreational harvest of luderick, by number, remained relatively constant between survey years. Interestingly, a comparison of the Spring seasons showed that the shore-based luderick harvest rate from the Southern Lake area had been significantly better during the first survey year. This was the only significant statistical difference detected for any seasonal harvest rate comparison involving luderick and may be explained by the exclusion of recreational fishers from the Eraring hot water outlet during the second survey year.

In summary, the harvest rate data have provided evidence of major changes in the Lake Macquarie fishery since the first survey year. These changes in seasonal harvest rates may be attributed to the effects of many inter-related factors, such as: (a) the availability of fish resulting from the removal of commercial fishing and/or natural fluctuations in abundance; (b) changes in targeting practices; and (c) increases in angler skill levels and technological improvements in fishing gear (e.g. the increased use of soft plastic lures may have led to increased harvest rates of dusky flathead).

Is there any evidence to indicate that the size of fish, crabs and cephalopods has changed since the first survey year? Changes in the size of fish can be assessed: (a) directly by comparing length frequency distributions, and their associated mean and median lengths; and (b) indirectly by comparing proportional changes in harvest levels (total number of individuals compared to total weight) between survey years. The change in size is inferred whenever the percentage change in harvest by number differs from the percentage change in harvest by weight. For example, when the percentage change by weight is greater than the percentage change by number, it can be inferred that the average size of fish has increased. Conversely, when the percentage change in harvest by weight is less than the percentage change by number, it can be inferred that the average size of fish has decreased.

An examination of comparative length frequency information, mean and median lengths between survey years indicated that most species were harvested at larger sizes during the post-RFH survey year. The mean and median sizes of dusky flathead, sand whiting, tailor, common squid, yellowfin bream, luderick, blue swimmer crab, large-toothed flounder and sand mullet were all larger during the second survey year (Figs. 24 to 34). The observed increase in mean size for dusky flathead was 6 cm , which was 3 cm more than the increase in minimum legal length that had been implemented since the first survey period. Interestingly, the increases in the mean and median size of common squid, blue swimmer crab and sand mullet occurred during the second survey year when their estimated harvests (number and weight) had decreased. Trumpeter whiting had identical mean and median lengths during each of the survey years. Yellow-finned leatherjacket taken during the second survey year had a smaller median length but identical mean length to fish taken during the first survey period (Fig.33).

Similar observations were made when comparing the relative changes in harvest (percentage number versus percentage weight) for these same species (Tables 3 and 4). Large increases in size during the second survey year were inferred for dusky flathead and tailor, moderate size increases were inferred for trumpeter whiting, sand whiting, yellowfin bream, luderick, blue swimmer crabs, common squid, large-toothed flounder and yellow-finned leatherjacket (Tables 3 and 4). No change in size was evident for sand mullet (Tables 3 and 4).

The removal of commercial fishing after the establishment of the RFH in 2002 meant that fish previously harvested by commercial fishers were now available to the recreational sector only. This management change may have led to an overall decrease in fishing pressure and a concomitant reduction in the rate of fishing mortality (commercial and recreational combined) on the fish, crab and squid stocks within Lake Macquarie. Any reduction in fishing effort or fishing mortality rate may allow the standing stocks of fish, crabs and squid some additional time to grow before they are harvested. If so, it would be expected that the mean and median sizes of many species should increase within the Lake Macquarie fishery. This is consistent with the increases in sizes observed during the post-RFH survey year.

## 6. CONCLUSIONS

This recreational fishing survey provides evidence of a relatively productive recreational fishery in Lake Macquarie and Swansea Channel. Comparisons made between two separate daytime recreational fishing surveys (the first done during the pre-RFH period and this second survey done during the post-RFH period) indicate that the post-RFH recreational fishery was very different to the fishery that existed prior to the implementation of the RFH. We documented statistically significant increases in recreational harvest for some prized recreational species and also some significant decreases for some other important recreational species. Overall, the indicators of recreational fishing quality that we examined indicated that the post-RFH fishery had improved in many ways since the pre-RFH survey period. A summary of the evidence provided in this report is that:
(a) the recreational harvest in both survey years was dominated by a relatively small number of taxa, however, the composition and relative contribution of these dominant taxa changed markedly between survey years. These changes occurred even though there was no significant difference between survey years in the total annual harvest, by number or weight, for the whole fishery;
(b) the recreational harvest of dusky flathead, tailor, sand whiting and trumpeter whiting (number and weight) and large-toothed flounder (weight only) had increased significantly during the postRFH survey year;
(c) the recreational harvest of common squid, yellow-finned leatherjacket and sand mullet, by number and weight, had decreased significantly during the post-RFH survey year;
(d) total fishing effort (boat and shore combined) showed little change (about 2.3\%), however, different trends were evident in the boat-based and shore-based fisheries. Fishing effort in the larger boat-based fishery increased by about $12.8 \%$ but this change was not statistically significant. In contrast, there was a statistically significant reduction of about $22.4 \%$ in the level of shore-based fishing.
(e) seasonal harvest rate comparisons between survey years tended to confirm the increasing or decreasing trends found in the annual recreational harvest estimates for the main species;
(f) comparisons of length frequency information, mean and median lengths between survey years indicated that most species were harvested at larger sizes during the post-RFH survey year. The mean and median sizes of dusky flathead, sand whiting, tailor, common squid, yellowfin bream, blue swimmer crab, large-toothed flounder and sand mullet were all larger during the second survey year.

## 7. RECOMMENDATIONS

1. This survey provides the first snapshot (point estimate) of the Lake Macquarie recreational fishery following the establishment of the waterway as a RFH. On-site surveys of recreational fishing are valuable tools for collecting information to describe the status of a fishery and any changes that may have occurred since previous survey periods. On-site surveys of the recreational fishery should be repeated regularly (every 3-5 years) to monitor the recreational fishery in Lake Macquarie.
2. It would be prudent and cost-effective to incorporate some biological sampling of key recreational species (e.g. age composition and reproductive biology) into any repeat survey work. Biological information will be invaluable for interpreting and understanding the factors that influence major changes in fish populations between survey periods.
3. Before future surveys or monitoring programmes are done in Lake Macquarie, it is recommended that statistical power analyses be done of the recreational fishing dataset collected during this study. Power analyses are vital for determining scientifically defensible and cost-effective survey designs that have sufficient statistical power to detect changes in the recreational fishery throughout time.

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## 9. APPENDICES

Appendix 1. Location of recognized boat ramps around Lake Macquarie.

| Ramp No. | Ramp Location |
| :---: | :---: |
| 1 | Blacksmiths - Ungala Street |
| 2 | Pelican - Lakeview Parade |
| 3 | Pelican - off Naru Street (near airport) |
| 4 | Belmont South - Paley Crescent |
| 5 | Belmont - Brooks Parade |
| 6 | Valetine - Bennett Park off Dilkera Avenue |
| 7 | Croudace Bay - Thomas Halton Park, Bareki Road |
| 8 | Eleebana - Lions Park, Bareki Road |
| 9 | Speers Point - Cockle Creek, Creek Reserve Road |
| 10 | Marmong Point - Off Nanda Street |
| 11 | Bolton Point - Off Middle Point Road |
| 12 | Toronto - Lions Park, Anzac Parade |
| 13 | Toronto - Wharf Street |
| 14 | Coal Point - Birraban Reserve Robey Crescent |
| 15 | Rathmines - Styles Point off Overhill Road |
| 16 | Rathmines - Rathmines Park off Dorrington Road |
| 17 | Balmoral - Letchworth Parade |
| 18 | Wangi Wangi - Kent Place |
| 19 | Wangi Wangi - Wangi Caravan Park (Watkins Road) |
| 20 | Wangi Wangi - off Dobell Drive (Wangi Wangi Beach) |
| 21 | Dora Creek - Dora Street |
| 22 | Bonnells Bay - Pendlebury Park, Grand Parade West |
| 23 | Balcolyn - Shingle Splitters Point King Street |
| 24 | Balcolyn - Balcolyn Street (near Progress Hall) |
| 25 | Sunshine - Sunshine Reserve off Sunshine Parade |
| 26 | Morisset Park - Lakeview Road |
| 27 | Wyee - Behind Mecca Caravan Park Ruttleys Road |
| 28 | Vales Point - off Peveril Street (The Cut) |
| 29 | Summerland Point - off Cams Boulevarde |
| 30 | Gwandalan - Garema Road |
| 31 | Gwandalan - off Koowong Road (Crangan Bay) |
| 32 | Gwandalan - off Gamban Road |
| 33 | Nords Wharf - Branter Road |
| 34 | Cams Wharf - Cams Wharf Road |
| 35 | Swansea - The Esplanade |
| 36 | Swansea - Fishermen's Co-op |
| 37 | Swansea - Coon Island 1 off Wallarah Street |
| 38 | Swansea - Coon Island 2 off Wallarah Street |

Appendix 2. The number of individuals observed ( N ), the number of individuals measured ( n ), size range ( cm ), median length ( cm ), and mean lengths ( cm ) for all taxa recorded with recreational fishers (boat and shore-based fishers combined) in the Lake Macquarie fishery during each survey year
TOTAL FISHERY

| COMMON NAME | SCIENTIFIC NAME | TOTAL FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Batfish, Silver | Monodactylus argenteus | 1 | 1 | 14 | 14.0 | 14.0 | 23 | 14 | 8 to 14 | 11.0 | 11.2 |
| Bigeye, Red | Priacanthus spp. | 13 | 13 | 25 to 28 | 26.0 | 26.4 | - | - | - | - |  |
| Blackfish, Rock | Girella elevata | 2 | 2 | 25 to 26 | 25.5 | 25.5 | 3 | 3 | 24 to 30 | 29.0 | 27.7 |
| Bream, Black | Acanthopagrus butcheri | 3 | 3 | 25 to 27 | 26.0 | 26.0 | - | - | - | - |  |
| Bream, Yellowfin | Acanthopagrus australis | 646 | 615 | 8 to 43 | 26.0 | 26.2 | 2546 | 2423 | 13 to 48 | 27.0 | 28.2 |
| Butterfish, Striped | Scatophagus argus | - | - | - | - | - | 6 | 6 | 20 to 31 | 21.5 | 23.2 |
| Calamari, Southern | Sepioteuthis australis | 11 | 7 | 23 to 36 | 28.0 | 28.9 | 81 | 73 | 8 to 45 | 30.0 | 28.4 |
| Cobia | Rachycentron canadum | 1 | 1 | 71 | 71.0 | 71.0 | 2 | 2 | 120 to 146 | 133.0 | 133.0 |
| Crab, Blue Swimmer | Portunus pelagicus | 1064 | 792 | 4 to 14 | 8.0 | 8.2 | 2516 | 2306 | 2 to 17 | 9.0 | 9.3 |
| Crab, Hairy-backed | Charybdis natator | 1 | - | - | - | - | - | - | - | - |  |
| Crab, Mud | Scylla serrata | 32 | 31 | 7 to 15 | 9.0 | 9.5 | 23 | 19 | 9 to 15 | 12.0 | 11.9 |
| Cuttlefish | Sepia spp. | - | - | - | - | - | 15 | 15 | 10 to 22 | 20.0 | 17.9 |
| Dolphin Fish | Coryphaena hippurus | 4 | 3 | 44 to 58 | 58.0 | 53.3 | - | - | - | - |  |
| Eel, Long-Finned | Anguilla reinhardtii | - | - | - | - | - | 1 | 1 | 44 | 44.0 | 44.0 |
| Eel, Short-Finned | Anguilla australis | - | - | - | - | - | 10 | 10 | 40 to 51 | 46.0 | 46.0 |
| Eels | Anguilliformes | 1 | 1 | 65 | 65.0 | 65.0 | - | - | - | - |  |
| Flathead, Dusky | Platycephalus fuscus | 322 | 311 | 20 to 80 | 42.0 | 43.7 | 1922 | 1867 | 22 to 98 | 48.0 | 49.9 |

Appendix 2, continued.

| COMMON NAME | SCIENTIFIC NAME | TOTAL FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Flathead, Eastern Blue-Spotted | Platycephalus caeruleopunctatus | 13 | 13 | 30 to 45 | 38.0 | 37.8 | 2 | 2 | 30 to 45 | 37.5 | 37.5 |
| Flathead, Long-Spined | Platycephalus longispinis | 1 | 1 | 24 | 24.0 | 24.0 | 1 | 1 | 30 | 30.0 | 30.0 |
| Flathead, Marbled | Platycephalus marmoratus | 1 | 1 | 51 | 51.0 | 51.0 | 6 | 5 | 40 to 68 | 52.0 | 52.4 |
| Flathead, Northern Sand | Platycephalus arenarius | 1 | 1 | 44 | 44.0 | 44.0 | 1 | 1 | 32 | 32.0 | 32.0 |
| Flathead, Tiger | Neoplatycephalus richardsoni | 1 | 1 | 68 | 68.0 | 68.0 | - | - | - | - | - |
| Flounder, Large-Toothed | Pseudorhombus arsius | 74 | 72 | 13 to 36 | 27.0 | 26.1 | 352 | 316 | 19 to 44 | 28.0 | 28.4 |
| Flounder, Small-Toothed | Pseudorhombus jenynsii | 73 | 66 | 12 to 40 | 26.0 | 26.1 | 144 | 139 | 10 to 41 | 27.0 | 27.6 |
| Garfish, River | Hyporhamphus regularis | 72 | 27 | 19 to 29 | 24.0 | 23.9 | 144 | 88 | 18 to 34 | 25.0 | 25.3 |
| Garfish, Sea | Hyporhamphus australis | 25 | 24 | 17 to 29 | 24.0 | 23.8 | 56 | 45 | 19 to 37 | 30.0 | 29.5 |
| Goatfish, Blue-Striped | Upeneichthys lineatus | 1 | - | - | - | - | - | - | - | - | - |
| Gurnard, Red | Chelidonichthys kumu | - | - | - | - | - | 3 | 3 | 30 to 35 | 32.0 | 32.3 |
| Herring, Giant | Elops machnata | 1 | 1 | 62 | 62.0 | 62.0 | - | - | - | - | - |
| Herring, Southern | Herklotsichthys castelnaui | 125 | 125 | 5 to 20 | 10.0 | 10.7 | 31 | 28 | 7 to 13 | 8.0 | 8.7 |
| Kingfish | Seriola lalandi | 3 | 3 | 23 to 76 | 55.0 | 51.3 | 14 | 13 | 36 to 80 | 62.0 | 60.0 |
| Leatherjacket, Chinaman | Nelusetta ayraudi | 5 | - | - | - | - | 23 | 23 | 18 to 48 | 24.0 | 26.0 |
| Leatherjacket, Fan-Bellied | Monacanthus chinensis | 78 | 63 | 10 to 32 | 25.0 | 24.2 | 36 | 35 | 20 to 45 | 30.0 | 30.0 |
| Leatherjacket, Rough | Scobinichthys granulatus | 7 | 7 | 20 to 32 | 26.0 | 25.3 | 12 | 12 | 12 to 37 | 33.0 | 31.1 |

Appendix 2, continued.

| COMMON NAME | SCIENTIFIC NAME | TOTAL FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Leatherjacket, Six-Spined | Meuschenia freycineti | 154 | 116 | 7 to 33 | 18.0 | 17.7 | 57 | 43 | 8 to 38 | 23.0 | 23.2 |
| Leatherjacket, Yellow-Finned | Meuschenia trachylepis | 249 | 173 | 8 to 32 | 24.0 | 22.4 | 170 | 138 | 10 to 40 | 20.5 | 21.7 |
| Longtom, Stout | Tylosurus gavialoides | 9 | 8 | 29 to 105 | 61.0 | 60.1 | 20 | 18 | 39 to 106 | 68.5 | 67.5 |
| Luderick | Girella tricuspidata | 1293 | 1226 | 22 to 44 | 28.0 | 28.5 | 2432 | 2136 | 19 to 46 | 30.0 | 30.6 |
| Mackerel, Frigate | Auxis thazard | - | - | - | - | - | 1 | 1 | 36 | 36.0 | 36.0 |
| Mackerel, Slimy | Scomber australasicus | - | - | - | - | - | 14 | 14 | 25 to 36 | 27.0 | 27.6 |
| Morwong, Red | Cheilodactylus fuscus | - | - | - | - | - | 1 | 1 | 28 | 28.0 | 28.0 |
| Mullet, Fan-Tail | Mugil georgii | 8 | 3 | 25 to 28 | 26.0 | 26.3 | 9 | 9 | 22 to 27 | 24.0 | 24.7 |
| Mullet, Flat-Tail | Liza argentea | 121 | 80 | 17 to 38 | 28.0 | 27.3 | 231 | 210 | 18 to 49 | 30.0 | 31.5 |
| Mullet, Sand | Myxus elongatus | 297 | 118 | 13 to 37 | 26.0 | 26.1 | 96 | 62 | 5 to 45 | 32.0 | 27.4 |
| Mullet, Sea | Mugil cephalus | 2 | 2 | 32 to 33 | 32.5 | 32.5 | 32 | 30 | 21 to 46 | 24.5 | 28.8 |
| Mulloway | Argyrosomus hololepidotus | 6 | 6 | 58 to 99 | 68.0 | 70.8 | 7 | 7 | 57 to 146 | 73.0 | 93.1 |
| Octopus | Octopus spp. | 1 | - | - | - | - | 30 | - | - | - |  |
| Perch, Butterfly | Caesioperca lepidoptera | - | - | - | - | - | 1 | 1 | 35 | 35.0 | 35.0 |
| Ray, Shovelnose | Rhinobatidae | 1 | 1 | 75 | 75.0 | 75.0 | 2 | 2 | 57 to 85 | 71.0 | 71.0 |
| Salmon, Australian | Arripis trutta | 1 | 1 | 58 | 58.0 | 58.0 | 18 | 18 | 42 to 80 | 55.0 | 58.2 |
| Scorpioncod, Red | Scorpaena cardinalis | - | - | - | - | - | 1 | 1 | 28 | 28.0 | 28.0 |

Appendix 2, continued.

| COMMON NAME | SCIENTIFIC NAME | TOTAL FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Seapike, Long-Finned | Dinolestes lewini | - | - | - | - | - | 6 | 6 | 32 to 44 | 33.5 | 35.8 |
| Seapike, Striped | Sphyraena obtusata | 1 | 1 | 30 | 30.0 | 30.0 | 35 | 35 | 20 to 50 | 30.0 | 33.6 |
| Shark, Wobbegong | Orectolobus spp. | - | - | - | - | - | 1 | 1 | 58 | 58.0 | 58.0 |
| Silver biddy | Gerres subfasciatus | - | - | - | - | - | 3 | 3 | 28 to 32 | 30.0 | 30.0 |
| Snapper | Pagrus auratus | 101 | 95 | 8 to 54 | 27.0 | 26.3 | 253 | 250 | 11 to 45 | 29.0 | 28.8 |
| Sole, Black | Synaptura nigra | 1 | - | - | - | - | - | - | - | - |  |
| Squid, Arrow | Nototodarus gouldi | 1 | 1 | 12 | 12.0 | 12.0 | - | - | - | - | - |
| Squid, Common | Photololigo spp. | 1296 | 919 | 6 to 50 | 16.0 | 16.6 | 1963 | 1502 | 6 to 40 | 18.0 | 17.8 |
| Stingarees \& Black Stingrays | Urolophidae \& Dasyatididae | - | - | - | - | - | 3 | - | - | - |  |
| Surgeon Fish | Acanthuridae | 1 | 1 | 27 | 27.0 | 27.0 | 1 | 1 | 34 | 34.0 | 34.0 |
| Sweep, Silver | Scorpis lineolatus | 1 | 1 | 15 | 15.0 | 15.0 | 10 | 10 | 16 to 32 | 18.0 | 19.6 |
| Tailor | Pomatomus saltatrix | 206 | 165 | 11 to 45 | 30.0 | 30.0 | 1544 | 1391 | 10 to 64 | 34.0 | 35.9 |
| Tarwhine | Rhabdosargus sarba | 71 | 61 | 18 to 31 | 24.0 | 23.9 | 123 | 109 | 17 to 46 | 26.0 | 25.6 |
| Teraglin | Atractoscion aequidens | 5 | 5 | 40 to 45 | 42.0 | 41.8 | - | - | - | - | - |
| Trevally, Black (Spinefoot) | Siganus spp. | 74 | 33 | 10 to 19 | 14.0 | 14.5 | 1 | 1 | 12 | 12.0 | 12.0 |
| Trevally, Silver | Pseudocaranx dentex | 8 | 8 | 23 to 59 | 25.0 | 29.1 | 53 | 53 | 17 to 59 | 29.0 | 32.0 |
| Trumpeter, Six-Lined | Pelates quadrilineatus | 14 | 10 | 6 to 10 | 7.5 | 7.7 | 12 | 4 | 8 to 12 | 9.5 | 9.8 |

Appendix 2, continued.

| COMMON NAME | SCIENTIFIC NAME | TOTAL FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Whiting, Sand | Sillago ciliata | 111 | 109 | 15 to 38 | 27.0 | 27.3 | 645 | 533 | 14 to 45 | 28.0 | 28.5 |
| Whiting, School | Sillago flindersi | 7 | 7 | 21 to 26 | 25.0 | 24.3 | 9 | 9 | 21 to 26 | 24.0 | 23.8 |
| Whiting, Trumpeter | Sillago maculata | 696 | 520 | 7 to 34 | 21.0 | 20.7 | 5541 | 4043 | 9 to 33 | 21.0 | 21.0 |
| Wirrah | Acanthistius ocellatus | 3 | 3 | 28 to 30 | 28.0 | 28.7 | - | - | - | - | - |
| Wrasse, Crimson-Banded | Notolabrus gymnogenis | 2 | 2 | 24 to 31 | 27.5 | 27.5 | 3 | 3 | 22 to 30 | 30.0 | 27.3 |
| Yellowtail | Trachurus novaezelandiae | 1 | 1 | 27 | 27.0 | 27.0 | 78 | 43 | 15 to 36 | 24.0 | 24.2 |

[^2]Appendix 3. The number of individuals observed ( N ), the number of individuals measured ( n ), size range ( cm ), median length ( cm ), and mean lengths ( cm ) for all taxa recorded with boat-based recreational fishers in the Lake Macquarie fishery during each survey year.

| COMMON NAME | SCIENTIFIC NAME | BOAT-BASED FISHERY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |  |  |
|  |  | N | n | Min | Max | Range | Median | Mean | N | n | Min | Max | Range | Median | Mean |
| Bigeye, Red | Priacanthus spp. | 13 | 13 | 25 | 28 | 25 to 28 | 26.0 | 26.4 | - | - |  |  | - | - | - |
| Blackfish, Rock | Girella elevata | - | - |  |  | - | - | - | 1 | 1 | 29 | 29 | 29 | 29.0 | 29.0 |
| Bream, Black | Acanthopagrus butcheri | 3 | 3 | 25 | 27 | 25 to 27 | 26.0 | 26.0 | - | - |  |  | - | - | - |
| Bream, Yellowfin | Acanthopagrus australis | 407 | 391 | 8 | 43 | 8 to 43 | 26.0 | 25.9 | 2131 | 2026 | 16 | 48 | 16 to 48 | 27.0 | 28.0 |
| Calamari, Southern | Sepioteuthis australis | 2 | 2 | 30 | 35 | 30 to 35 | 32.5 | 32.5 | 62 | 59 | 8 | 45 | 8 to 45 | 30.0 | 29.2 |
| Cobia | Rachycentron canadum | 1 | 1 | 71 | 71 | 71 | 71.0 | 71.0 | 2 | 2 | 120 | 146 | 120 to 146 | 133.0 | 133.0 |
| Crab, Blue Swimmer | Portunus pelagicus | 1039 | 784 | 4 | 14 | 4 to 14 | 8.0 | 8.3 | 2476 | 2266 | 2 | 17 | 2 to 17 | 9.0 | 9.3 |
| Crab, Hairy-backed | Charybdis natator | 1 | - | 0 | 0 | - | - | - | - | - |  |  | - | - | - |
| Crab, Mud | Scylla serrata | 32 | 31 | 7 | 15 | 7 to 15 | 9.0 | 9.5 | 20 | 16 | 9 | 15 | 9 to 15 | 11.0 | 11.6 |
| Cuttlefish | Sepia spp. | - | - |  |  | - | - | - | 12 | 12 | 16 | 22 | 16 to 22 | 20.0 | 19.3 |
| Dolphin Fish | Coryphaena hippurus | 4 | 3 | 44 | 58 | 44 to 58 | 58.0 | 53.3 | - | - |  |  | - | - | - |
| Eel, Short-Finned | Anguilla australis | - | - |  |  | - | - | - | 10 | 10 | 40 | 51 | 40 to 51 | 46.0 | 46.0 |
| Flathead, Dusky | Platycephalus fuscus | 289 | 278 | 20 | 80 | 20 to 80 | 41.0 | 43.2 | 1733 | 1681 | 22 | 98 | 22 to 98 | 48.0 | 50.7 |
| Flathead, Eastern Blue-Spotted | Platycephalus caeruleopunctatus | 13 | 13 | 30 | 45 | 30 to 45 | 38.0 | 37.8 | 2 | 2 | 30 | 45 | 30 to 45 | 37.5 | 37.5 |
| Flathead, Long-Spined | Platycephalus longispinis | 1 | 1 | 24 | 24 | 24 | 24.0 | 24.0 | 1 | 1 | 30 | 30 | 30 | 30.0 | 30.0 |
| Flathead, Marbled | Platycephalus marmoratus | - | - |  |  | - | - | - | 6 | 5 | 40 | 68 | 40 to 68 | 52.0 | 52.4 |
| Flathead, Northern Sand | Platycephalus arenarius | 1 | 1 | 44 | 44 | 44 | 44.0 | 44.0 | 1 | 1 | 32 | 32 | 32 | 32.0 | 32.0 |

Appendix 3, continued.

| COMMON NAME | SCIENTIFIC NAME | BOAT-BASED FISHERY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |  |  |
|  |  | N | n | Min | Max | Range | Median | Mean | N | n | Min | Max | Range | Median | Mean |
| Flounder, Large-Toothed | Pseudorhombus arsius | 73 | 71 | 13 | 36 | 13 to 36 | 27.0 | 26.1 | 332 | 296 | 19 | 44 | 19 to 44 | 28.0 | 28.4 |
| Flounder, Small-Toothed | Pseudorhombus jenynsii | 72 | 65 | 12 | 40 | 12 to 40 | 26.0 | 26.1 | 139 | 134 | 10 | 41 | 10 to 41 | 27.0 | 27.6 |
| Garfish, River | Hyporhamphus regularis | 72 | 27 | 19 | 29 | 19 to 29 | 24.0 | 23.9 | 84 | 46 | 20 | 30 | 20 to 30 | 25.0 | 24.7 |
| Garfish, Sea | Hyporhamphus australis | 25 | 24 | 17 | 29 | 17 to 29 | 24.0 | 23.8 | 46 | 35 | 19 | 37 | 19 to 37 | 28.0 | 28.8 |
| Goatfish, Blue-Striped | Upeneichthys lineatus | 1 | - | 0 | 0 | - | - | - | - | - |  |  | - | - | - |
| Gurnard, Red | Chelidonichthys kumu | - | - |  |  | - | - | - | 3 | 3 | 30 | 35 | 30 to 35 | 32.0 | 32.3 |
| Kingfish | Seriola lalandi | 2 | 2 | 23 | 55 | 23 to 55 | 39.0 | 39.0 | 11 | 10 | 36 | 80 | 36 to 80 | 56.0 | 58.3 |
| Leatherjacket, Chinaman | Nelusetta ayraudi | - | - |  |  | - | - | - | 20 | 20 | 18 | 48 | 18 to 48 | 24.5 | 26.6 |
| Leatherjacket, Fan-Bellied | Monacanthus chinensis | 56 | 46 | 10 | 32 | 10 to 32 | 25.0 | 24.2 | 32 | 31 | 20 | 45 | 20 to 45 | 32.0 | 31.0 |
| Leatherjacket, Rough | Scobinichthys granulatus | 6 | 6 | 20 | 32 | 20 to 32 | 26.0 | 25.8 | 9 | 9 | 27 | 37 | 27 to 37 | 35.0 | 33.1 |
| Leatherjacket, Six-Spined | Meuschenia freycineti | 45 | 16 | 11 | 33 | 11 to 33 | 21.5 | 20.9 | 18 | 17 | 21 | 33 | 21 to 33 | 29.0 | 28.5 |
| Leatherjacket, Yellow-Finned | Meuschenia trachylepis | 179 | 108 | 9 | 32 | 9 to 32 | 22.5 | 22.0 | 125 | 93 | 10 | 40 | 10 to 40 | 18.0 | 20.8 |
| Longtom, Stout | Tylosurus gavialoides | 3 | 2 | 65 | 72 | 65 to 72 | 68.5 | 68.5 | 15 | 14 | 40 | 106 | 40 to 106 | 68.5 | 70.6 |
| Luderick | Girella tricuspidata | 86 | 67 | 24 | 37 | 24 to 37 | 28.0 | 28.7 | 371 | 280 | 23 | 40 | 23 to 40 | 31.0 | 30.7 |
| Mackerel, Frigate | Auxis thazard | - | - |  |  | - | - | - | 1 | 1 | 36 | 36 | 36 | 36.0 | 36.0 |
| Mackerel, Slimy | Scomber australasicus | - | - |  |  | - | - | - | 14 | 14 | 25 | 36 | 25 to 36 | 27.0 | 27.6 |
| Mullet, Fan-Tail | Mugil georgii | - | - |  |  | - | - | - | 7 | 7 | 23 | 27 | 23 to 27 | 25.0 | 25.3 |

Appendix 3, continued.

| COMMON NAME | SCIENTIFIC NAME | BOAT-BASED FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Mullet, Flat-Tail | Liza argentea | 71 | 37 | 20 to 37 | 28.0 | 29.3 | 189 | 177 | 18 to 45 | 30.0 | 31.3 |
| Mullet, Sand | Myxus elongatus | 151 | 64 | 18 to 37 | 26.0 | 26.3 | 26 | 26 | 27 to 45 | 37.5 | 37.1 |
| Mullet, Sea | Mugil cephalus | - | - | - | - | - | 5 | 4 | 36 to 46 | 43.0 | 42.0 |
| Mulloway | Argyrosomus hololepidotus | 5 | 5 | 58 to 99 | 67.0 | 70.6 | 7 | 7 | 57 to 146 | 73.0 | 93.1 |
| Octopus | Octopus spp. | 1 | - | - | - | - | 26 | - | - | - | - |
| Perch, Butterfly | Caesioperca lepidoptera | - | - | - | - | - | 1 | 1 | 35 | 35.0 | 35.0 |
| Ray, Shovelnose | Rhinobatidae | 1 | 1 | 75 | 75.0 | 75.0 | 2 | 2 | 57 to 85 | 71.0 | 71.0 |
| Salmon, Australian | Arripis trutta | - | - | - | - | - | 18 | 18 | 42 to 80 | 55.0 | 58.2 |
| Scorpioncod, Red | Scorpaena cardinalis | - | - | - | - | - | 1 | 1 | 28 | 28.0 | 28.0 |
| Seapike, Long-Finned | Dinolestes lewini | - | - | - | - | - | 2 | 2 | 40 to 44 | 42.0 | 42.0 |
| Seapike, Striped | Sphyraena obtusata | 1 | 1 | 30 | 30.0 | 30.0 | 34 | 34 | 20 to 50 | 30.0 | 33.3 |
| Shark, Wobbegong | Orectolobus spp. | - | - | - | - | - | 1 | 1 | 58 | 58.0 | 58.0 |
| Silver biddy | Gerres subfasciatus | - | - | - | - | - | 3 | 3 | 28 to 32 | 30.0 | 30.0 |
| Snapper | Pagrus auratus | 97 | 91 | 10 to 54 | 27.0 | 26.8 | 247 | 244 | 19 to 45 | 29.0 | 28.9 |
| Sole, Black | Synaptura nigra | 1 | - | - | - | - | - | - | - | - | - |
| Squid, Common | Photololigo spp. | 1200 | 833 | 6 to 33 | 16.0 | 16.7 | 1855 | 1416 | 6 to 40 | 18.0 | 17.8 |
| Stingarees \& Black Stingrays | Urolophidae \& Dasyatididae | - | - | - | - | - | 3 | - | - | - | - |

Appendix 3, continued.

| COMMON NAME | SCIENTIFIC NAME | BOAT-BASED FISHERY |  |  |  |  |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |  |  |
|  |  | N | n | Min | Max | Range | Median | Mean | N | n | Min | Max | Range | Median | Mean |
| Surgeon Fish | Acanthuridae | 1 | 1 | 27 | 27 | 27 | 27.0 | 27.0 | - | - |  |  | - | - | - |
| Sweep, Silver | Scorpis lineolatus | 1 | 1 | 15 | 15 | 15 | 15.0 | 15.0 | 2 | 2 | 26 | 32 | 26 to 32 | 29.0 | 29.0 |
| Tailor | Pomatomus saltatrix | 176 | 135 | 11 | 45 | 11 to 45 | 32.0 | 31.3 | 1428 | 1279 | 10 | 64 | 10 to 64 | 34.0 | 36.4 |
| Tarwhine | Rhabdosargus sarba | 6 | 2 | 22 | 30 | 22 to 30 | 26.0 | 26.0 | 84 | 80 | 18 | 46 | 18 to 46 | 26.0 | 25.4 |
| Teraglin | Atractoscion aequidens | 5 | 5 | 40 | 45 | 40 to 45 | 42.0 | 41.8 | - | - |  |  | - | - |  |
| Trevally, Silver | Pseudocaranx dentex | - | - |  |  | - | - | - | 43 | 43 | 17 | 50 | 17 to 50 | 28.0 | 29.7 |
| Trumpeter, Six-Lined | Pelates quadrilineatus | 2 | - | 0 | 0 | - | - | - | 8 | - | 0 | 0 | - | - |  |
| Whiting, Sand | Sillago ciliata | 96 | 94 | 19 | 38 | 19 to 38 | 27.5 | 27.9 | 593 | 481 | 14 | 45 | 14 to 45 | 28.0 | 28.6 |
| Whiting, School | Sillago flindersi | 7 | 7 | 21 | 26 | 21 to 26 | 25.0 | 24.3 | 9 | 9 | 21 | 26 | 21 to 26 | 24.0 | 23.8 |
| Whiting, Trumpeter | Sillago maculata | 674 | 498 | 7 | 34 | 7 to 34 | 21.0 | 20.7 | 4975 | 3493 | 9 | 33 | 9 to 33 | 21.0 | 21.2 |
| Wirrah | Acanthistius ocellatus | 3 | 3 | 28 | 30 | 28 to 30 | 28.0 | 28.7 | - | - |  |  | - | - | - |
| Wrasse, Crimson-Banded | Notolabrus gymnogenis | 1 | 1 | 31 | 31 | 31 | 31.0 | 31.0 | 2 | 2 | 30 | 30 | 30 | 30.0 | 30.0 |
| Yellowtail | Trachurus novaezelandiae | 1 | 1 | 27 | 27 | 27 | 27.0 | 27.0 | 75 | 40 | 19 | 36 | 19 to 36 | 24.0 | 24.9 |

[^3]Appendix 4. The number of individuals observed ( N ), the number of individuals measured ( n ), size range ( cm ), median length ( cm ), and mean lengths ( cm )

| COMMON NAME | SCIENTIFIC NAME | SHORE-BASED FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Batfish, Silver | Monodactylus argenteus | 1 | 1 | 14 | 14.0 | 14.0 | 23 | 14 | 8 to 14 | 11.0 | 11.2 |
| Blackfish, Rock | Girella elevata | 2 | 2 | 25 to 26 | 25.5 | 25.5 | 2 | 2 | 24 to 30 | 27.0 | 27.0 |
| Bream, Yellowfin | Acanthopagrus australis | 239 | 224 | 9 to 42 | 27.0 | 26.6 | 415 | 397 | 13 to 45 | 29.0 | 29.0 |
| Butterfish, Striped | Scatophagus argus | - | - | - | - | - | 6 | 6 | 20 to 31 | 21.5 | 23.2 |
| Calamari, Southern | Sepioteuthis australis | 9 | 5 | 23 to 36 | 26.0 | 27.4 | 19 | 14 | 15 to 35 | 24.0 | 24.9 |
| Crab, Blue Swimmer | Portunus pelagicus | 25 | 8 | 7 to 9 | 8.0 | 7.8 | 40 | 40 | 6 to 16 | 9.0 | 9.6 |
| Crab, Mud | Scylla serrata | - | - | - | - | - | 3 | 3 | 13 to 14 | 14.0 | 13.7 |
| Cuttlefish | Sepia spp. | - | - | - | - | - | 3 | 3 | 10 to 15 | 12.0 | 12.3 |
| Eel, Long-Finned | Anguilla reinhardtii | - | - | - | - | - | 1 | 1 | 44 | 44.0 | 44.0 |
| Eels | Anguilliformes | 1 | 1 | 65 | 65.0 | 65.0 | - | - | - | - | - |
| Flathead, Dusky | Platycephalus fuscus | 33 | 33 | 24 to 72 | 47.0 | 48.2 | 189 | 186 | 23 to 80 | 41.0 | 43.3 |
| Flathead, Marbled | Platycephalus marmoratus | 1 | 1 | 51 | 51.0 | 51.0 | - | - | - | - | - |
| Flathead, Tiger | Neoplatycephalus richardsoni | 1 | 1 | 68 | 68.0 | 68.0 | - | - | - | - | - |
| Flounder, Large-Toothed | Pseudorhombus arsius | 1 | 1 | 27 | 27.0 | 27.0 | 20 | 20 | 20 to 40 | 28.0 | 28.6 |
| Flounder, Small-Toothed | Pseudorhombus jenynsii | 1 | 1 | 28 | 28.0 | 28.0 | 5 | 5 | 18 to 33 | 30.0 | 27.0 |
| Garfish, River | Hyporhamphus regularis | - | - | - | - | - | 60 | 42 | 18 to 34 | 26.0 | 25.8 |
| Garfish, Sea | Hyporhamphus australis | - | - | - | - | - | 10 | 10 | 25 to 35 | 32.5 | 31.8 |

Appendix 4, continued.

| COMMON NAME | SCIENTIFIC NAME | SHORE-BASED FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Herring, Giant | Elops machnata | 1 | 1 | 62 | 62.0 | 62.0 | - | - | - | - |  |
| Herring, Southern | Herklotsichthys castelnaui | 125 | 125 | 5 to 20 | 10.0 | 10.7 | 31 | 28 | 7 to 13 | 8.0 | 8.7 |
| Kingfish | Seriola lalandi | 1 | 1 | 76 | 76.0 | 76.0 | 3 | 3 | 63 to 71 | 63.0 | 65.7 |
| Leatherjacket, Chinaman | Nelusetta ayraudi | 5 | - | - | - | - | 3 | 3 | 20 to 24 | 21.0 | 21.7 |
| Leatherjacket, Fan-Bellied | Monacanthus chinensis | 22 | 17 | 16 to 31 | 25.0 | 24.1 | 4 | 4 | 20 to 25 | 23.0 | 22.8 |
| Leatherjacket, Rough | Scobinichthys granulatus | 1 | 1 | 22 | 22.0 | 22.0 | 3 | 3 | 12 to 33 | 30.0 | 25.0 |
| Leatherjacket, Six-Spined | Meuschenia freycineti | 109 | 100 | 7 to 32 | 17.0 | 17.2 | 39 | 26 | 8 to 38 | 20.0 | 19.7 |
| Leatherjacket, Yellow-Finned | Meuschenia trachylepis | 70 | 65 | 8 to 30 | 24.0 | 23.0 | 45 | 45 | 14 to 36 | 22.0 | 23.4 |
| Longtom, Stout | Tylosurus gavialoides | 6 | 6 | 29 to 105 | 54.0 | 57.3 | 5 | 4 | 39 to 72 | 57.5 | 56.5 |
| Luderick | Girella tricuspidata | 1207 | 1159 | 22 to 44 | 28.0 | 28.5 | 2061 | 1856 | 19 to 46 | 30.0 | 30.6 |
| Morwong, Red | Cheilodactylus fuscus | - | - | - | - | - | 1 | 1 | 28 | 28.0 | 28.0 |
| Mullet, Fan-Tail | Mugil georgii | 8 | 3 | 25 to 28 | 26.0 | 26.3 | 2 | 2 | 22 to 23 | 22.5 | 22.5 |
| Mullet, Flat-Tail | Liza argentea | 50 | 43 | 17 to 38 | 27.0 | 25.5 | 42 | 33 | 27 to 49 | 31.0 | 32.8 |
| Mullet, Sand | Myxus elongatus | 146 | 54 | 13 to 36 | 27.0 | 25.9 | 70 | 36 | 5 to 39 | 16.5 | 20.3 |
| Mullet, Sea | Mugil cephalus | 2 | 2 | 32 to 33 | 32.5 | 32.5 | 27 | 26 | 21 to 43 | 23.0 | 26.7 |
| Mulloway | Argyrosomus hololepidotus | 1 | 1 | 72 | 72.0 | 72.0 | - | - | - | - |  |
| Octopus | Octopus spp. | - | - | - | - | - | 4 | - | - | - |  |

Appendix 4, continued.

| COMMON NAME | SCIENTIFIC NAME | SHORE-BASED FISHERY |  |  |  |  |  |  |  |  |  |
| :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: | :---: |
|  |  | SURVEY YEAR 1 <br> (March 1999 to February 2000) |  |  |  |  | SURVEY YEAR 2 <br> (December 2003 to November 2004) |  |  |  |  |
|  |  | N | n | Range | Median | Mean | N | n | Range | Median | Mean |
| Salmon, Australian | Arripis trutta | 1 | 1 | 58 | 58.0 | 58.0 | - | - | - | - | - |
| Seapike, Long-Finned | Dinolestes lewini | - | - | - | - | - | 4 | 4 | 32 to 34 | 32.5 | 32.8 |
| Seapike, Striped | Sphyraena obtusata | - | - | - | - | - | 1 | 1 | 45 | 45.0 | 45.0 |
| Snapper | Pagrus auratus | 4 | 4 | 8 to 18 | 15.0 | 14.0 | 6 | 6 | 11 to 30 | 26.0 | 24.0 |
| Squid, Arrow | Nototodarus gouldi | 1 | 1 | 12 | 12.0 | 12.0 | - | - | - | - | - |
| Squid, Common | Photololigo spp. | 96 | 86 | 6 to 50 | 14.0 | 15.2 | 108 | 86 | 9 to 33 | 18.0 | 18.0 |
| Surgeon Fish | Acanthuridae | - | - | - | - | - | 1 | 1 | 34 | 34.0 | 34.0 |
| Sweep, Silver | Scorpis lineolatus | - | - | - | - | - | 8 | 8 | 16 to 18 | 17.5 | 17.3 |
| Tailor | Pomatomus saltatrix | 30 | 30 | 18 to 32 | 23.5 | 23.9 | 116 | 112 | 10 to 56 | 32.0 | 30.4 |
| Tarwhine | Rhabdosargus sarba | 65 | 59 | 18 to 31 | 24.0 | 23.8 | 39 | 29 | 17 to 34 | 27.0 | 26.3 |
| Trevally, Black (Spinefoot) | Siganus spp. | 74 | 33 | 10 to 19 | 14.0 | 14.5 | 1 | 1 | 12 | 12.0 | 12.0 |
| Trevally, Silver | Pseudocaranx dentex | 8 | 8 | 23 to 59 | 25.0 | 29.1 | 10 | 10 | 28 to 59 | 40.5 | 41.9 |
| Trumpeter, Six-Lined | Pelates quadrilineatus | 12 | 10 | 6 to 10 | 7.5 | 7.7 | 4 | 4 | 8 to 12 | 9.5 | 9.8 |
| Whiting, Sand | Sillago ciliata | 15 | 15 | 15 to 30 | 25.0 | 24.1 | 52 | 52 | 15 to 40 | 27.5 | 27.8 |
| Whiting, Trumpeter | Sillago maculata | 22 | 22 | 8 to 26 | 22.0 | 21.1 | 566 | 550 | 10 to 30 | 20.0 | 20.0 |
| Wrasse, Crimson-Banded | Notolabrus gymnogenis | 1 | 1 | 24 | 24.0 | 24.0 | 1 | 1 | 22 | 22.0 | 22.0 |
| Yellowtail | Trachurus novaezelandiae | - | - | - | - | - | 3 | 3 | 15 | 15.0 | 15.0 |

- no observations or measurements made


## 10. SURVEY PERSONNEL

The following tables present lists of persons who worked in either the 1999/2000 and/or the 2003/2004 survey as field staff. Persons are listed according to their affiliation with an organised group. We again thank all of the following personnel for their valuable contributions to this project.

Survey personnel from the Lake Macquarie Concerned Anglers Group and NSW DPI Fishcare Volunteer program.

| Name | Affiliation | $\begin{gathered} \text { Survey Year } 1 \\ 1999 / 2000 \end{gathered}$ | $\begin{gathered} \text { Survey Year } 2 \\ 2003 / 2004 \end{gathered}$ |
| :---: | :---: | :---: | :---: |
| Steve Anderson | 1 | * |  |
| Tom Archbold | 1 | * |  |
| Brian Arnold | 2 |  | * |
| Col Austin | 1 | * |  |
| Liz Bailey | 1 | * |  |
| Ian Beresford | 1 |  | * |
| Winstone Buffrey | 1 | * |  |
| George Burrell | 1 |  | * |
| John Cheyne | 2 |  | * |
| Graham Clark | 2 |  | * |
| Frank Druery | 1 | * |  |
| Bill Gray | 1 | * |  |
| Patricia Hall | 1 | * |  |
| Richard Hall | 1 | * |  |
| Graham Halley | 1 | * |  |
| Ron Hemsley | 1 | * |  |
| Brian Hilton | 2 |  | * |
| Jack Howell | 1 |  | * |
| Angelo Iacono | 2 |  | * |
| Lionel Jones | 1 | * | * |
| Craig Jones | 1 | * |  |
| Lionel Jones | 1 | * |  |

Survey personnel from the Lake Macquarie Concerned Anglers Group and NSW DPI Fishcare Volunteer program, continued.

| Name | Affiliation | $\begin{gathered} \text { Survey Year } 1 \\ 1999 / 2000 \end{gathered}$ | Survey Year 2 2003/2004 |
| :---: | :---: | :---: | :---: |
| Alan Keft | 1 | * |  |
| John Lightfoot | 1,2 | * | * |
| Jan McLeod | 1,2 | * | * |
| John McLeod | 1,2 | * | * |
| Dennis Morgan | 1 | * |  |
| Col Munro | 1 | * | * |
| Gary Pearce | 1,2 | * | * |
| Allen Rae | 1 | * |  |
| Ray Searle | 2 |  | * |
| Noel Stoops | 1 | * | * |
| Rosalind Stoops | 1 |  | * |
| Kevin Turner | 1 | * |  |
| Noel Vidler | 1 | * |  |
| Ray Ward | 1 | * |  |
| Sam Wilson | 1,2 | * | * |
| Gail Young | 1 |  | * |

## Affiliation Key:

1 - Lake Macquarie Concerned Anglers Group
2- NSW DPI Fishcare Volunteers

* denotes participation in either survey year 1 and/or survey year 2

Survey personnel affiliated with B\&L Fishing and Cruises.

| Name | $\begin{gathered} \hline \text { Survey Year } 1 \\ 1999 / 2000 \\ \hline \end{gathered}$ | $\begin{gathered} \hline \text { Survey Year } 2 \\ 2003 / 2004 \\ \hline \end{gathered}$ |
| :---: | :---: | :---: |
| Steve Baggs |  | * |
| Jason Bennett | * |  |
| Michael Chipchase |  | * |
| Brett Corbett | * | * |
| Adrian Cornwall |  | * |
| Ted Doggett |  | * |
| Michael Ede |  | * |
| Aaron Edwards |  | * |
| Dane Haigh |  | * |
| Elise Harris |  | * |
| Dave Harris | * |  |
| Chris Hird | * |  |
| Ian Hobden |  | * |
| Ian Hobden | * |  |
| Ben Howe |  | * |
| Steven Jest |  | * |
| Clyde Kelton |  | * |
| Aaron Kinghorn | * |  |
| Eric McGilvray |  | * |
| Tim McGowan | * |  |
| Brad Minors |  | * |
| Steve Minors | * | * |
| Fiona Minors | * | * |
| Ben Minors | * |  |
| Linda Minors | * |  |
| Norma Minors | * |  |
| Keith Morgan |  | * |
| Micheal Mottley |  | * |
| Will Paul |  | * |
| Merilee Prangell |  | * |
| Garth Quick |  | * |
| Sharon Simington |  | * |
| Ryan Spong | * |  |

[^4]
## Survey personnel affiliated with the Swansea Australian Volunteer Coast Guard.

| Name | Survey Year 1 <br> $\mathbf{1 9 9 9} / \mathbf{2 0 0 0}$ | Survey Year 2 <br> $\mathbf{2 0 0 3} / \mathbf{2 0 0 4}$ |
| :--- | :---: | :---: |
| Larry Baker | $*$ |  |
| Nola Ellis | $*$ |  |
| Norm Ellis | $*$ |  |
| Allan Ferrier | $*$ |  |
| John Green | $*$ |  |
| Mike Hawkins | $*$ |  |
| Garry Horgan | $*$ |  |
| Tony Mackay | $*$ |  |
| John Margarie | $*$ |  |
| Craig Mason | $*$ |  |
| T. Milton | $*$ |  |
| Bruce Oliver | $*$ |  |
| John Rains | $*$ |  |
| Kay Rains | $*$ |  |
| Bill Rogan | $*$ |  |
| Barry Shoesmith | $*$ |  |
| Peter Skinner | $*$ |  |
| Bob Suttie | $*$ |  |
| Jan Suttie | $*$ |  |
| Aachard Taubman Thomas | $*$ |  |
| Doug Young | $*$ |  |

* denotes participation in either survey year 1 and/or survey year 2


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[^0]:    Key:
    \# Expanded estimates of harvest have not been calculated. This observation was classified as a rare event during this time period and its occurrence is simply noted. Not recorded or not calculated for rare event occurrences.

    Associated estimates of expanded weight $(\mathrm{kg})$ are not provided for this taxon in Tables 6 and 7 because a suitable length to weight conversion key was not available Other taxa details are provided in Appendix 3.
    $\begin{array}{ll}\text { * } & \text { Significantly different, } \mathrm{p}<0.05 \text {. } \\ \text { ns } & \text { No significant difference, } \mathrm{p}>0.05\end{array}$

[^1]:    Figure 27. Yellowfin bream - annual length frequency distributions and comparison between survey years for the whole recreational fishery in Lake

[^2]:    - no observations or measurements made

[^3]:    - no observations or measurements made

[^4]:    * denotes participation in either survey year 1 and/or survey year 2

